RESEARCH ARTICLE

Rethinking the nature of intraspecific variability and its consequences on species coexistence

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Abstract

Intraspecific variability (IV) has been proposed to explain species coexistence in diverse communities. Assuming, sometimes implicitly, that conspecific individuals can perform differently in the same environment and that IV increases niche overlap, previous studies have found contrasting results regarding the effect of IV on species coexistence. We aim at showing that the large IV observed in data does not mean that conspecific individuals are necessarily different in their response to the environment and that the role of high-dimensional environmental variation in determining IV has largely remained unexplored in forest plant communities. We first used a simulation experiment where an individual attribute is derived from a high-dimensional model, representing "perfect knowledge" of individual response to the environment, to illustrate how large observed IV can result from "imperfect knowledge" of the environment. Second, using growth data from clonal *Eucalyptus* plantations in Brazil, we estimated a major contribution of the environment in determining individual growth. Third, using tree growth data from long-term tropical forest inventories in French Guiana, Panama and India, we showed that tree growth in tropical forests is structured spatially and that despite a large observed IV at the population level, conspecific individuals perform more similarly locally than compared with heterospecific individuals. As the number of environmental dimensions that are well quantified at fine scale is generally lower than the actual number of dimensions influencing individual attributes, a great part of observed IV might be represented as random variation across individuals when in fact it is environmentally driven. This mis-representation has important consequences for inference about community dynamics. We emphasize that observed IV does not necessarily impact species coexistence per se but can reveal species response to high-dimensional environment, which is consistent with niche theory and the observation of the many differences between species in nature.

1 | INTRODUCTION

Ecological communities are characterized by numerous coexisting species, for instance in grasslands, coral reefs, or tropical forests. Understanding how these species coexist while competing for the same basic resources, viz. light, water, and nutrients (Baraloto et al., 2010), is a long-standing question in ecology (Gause, 1934; Hutchinson, 1961; Levine et al., 2017). Although numerous mechanisms have been suggested to contribute to species coexistence (Chesson, 2000; Connell, 1971; Hubbell, 2001; Janzen, 1970; Levine & HilleRisLambers, 2009; Wright, 2002), it is unclear when and to what extent they explain the high species diversity observed in nature (Clark, 2010). This is especially true in forests, where tree species coexist while seemingly requiring similar resources in the same location. Astonishingly, a hectare of tropical forest can harbor more than 900 plant species of a diversity of forms and functions (Wilson et al., 2012).

Many theoretical mechanisms that might explain tree species coexistence were explored using the underlying assumption, explicit or not, that all conspecific individuals are identical (e.g., Lotka (1925) and Volterra's (1926) competition models, Tilman's R*, Pacala et al., 1996; Rees, 2001). However, intraspecific variability (IV) in traits, demographic rates or any proxy of performance, henceforth denoted as "attributes," can alter community structure and dynamics (Bolnick et al., 2011). Indeed, large IV has been observed across a number of attributes in plant communities (Albert et al., 2012; Violle et al., 2012). For instance, Siefert et al. (2015) estimated that IV accounted for 25% of the variability in functional traits within plant communities on average, and this proportion was even estimated at 44% in a tropical forest (Poorter et al., 2018). Likewise, IV in growth rates for trees of standardized size, local crowding, terrain slope, and annual effect has been found to account for up to 38% of total growth variability in a tropical forest stand (Le Bec et al., 2015).

IV, as a pathway for coexistence, has so far not shared the same attention as other mechanisms. This is in part because modeling studies that have explored the effect of IV on species coexistence have yielded contrasting results (Stump et al., 2021). In most theoretical analyses, variability in attributes among conspecific individuals has been included through independent random draws (Barabás & D'Andrea, 2016; Crawford et al., 2019; Hart et al., 2016; Lichstein et al., 2007; but see Banitz, 2019; Purves & Vanderwel, 2014). Similarly, empirical studies typically summarize IV as a variance around species mean attributes (Albert et al., 2010; Jung et al., 2010; Poorter et al., 2018; Siefert et al., 2015). With this representation, IV can increase species niche overlap, thus making species functionally less different (i.e., "blurring species differences,"

Hart et al., 2016; Lichstein et al., 2007), sometimes slowing down competitive exclusion in models of community dynamics (Crawford et al., 2019; Vieilledent et al., 2010). However, in some other models, nonlinear responses can make such IV beneficial to the superior competitors (i.e., the most competitive individuals of the more competitive species), thus accelerating competitive exclusion (e.g., Courbaud et al., 2012; Hart et al., 2016). Alternatively, in specific spatial configurations, more precisely when IV is greater in species preferred habitats, it has been shown to foster species coexistence (Uriarte & Menge, 2018). Stump et al. (2021) have proposed to reconcile these contrasting results by distinguishing the effect of IV on niche traits (which control individual performance response to environmental conditions) versus hierarchical traits (which control individual performance independently from environmental conditions). They demonstrated with different simulation models of community dynamics that IV in traits can alter stabilizing mechanisms and fitness differences in a complex way which depends upon the nature of the traits (niche vs. hierarchical) and their response curve, and thus promote or not species coexistence. In all the above examples however, IV, since simulated through independent random draws around species mean attributes, would be caused by differences among individuals that are fully independent of the environment: differences among individuals would remain unchanged even when experiencing exactly the same environmental conditions. Importantly, such simulated IV thus leads to a variation among conspecific individuals that is completely unstructured in space and time. New appreciation of fine-scale environmental heterogeneity and structure as well as species differences in their response to the environment, however, may suggest that this assumption of unstructured IV is rarely met.

Novel remote sensing tools such as high-spatial and -temporal resolution airborne LiDAR scans (Cushman et al., 2022; Tymen et al., 2017), intensive soil samplings and metabarcoding (Zinger et al., 2019), and more generally studies on the microclimate (Zellweger et al., 2019) and microhabitats (Baraloto & Couteron, 2010) have indeed evidenced strong environmental variation operating at fine scales (e.g., cm to meter scales) in many dimensions (Figure 1). These environmental dimensions can not only be resources for which species compete (e.g., light, water, nutrients) but also for all other components that shape the environment locally in space and time (e.g., temperature, wind, elevation, slope, soil texture, soil microorganisms etc.). In parallel, naturalists and taxonomists have long documented species differences in many aspects of their morphology and life history (Figure 2). Such differences between species have then been specified and quantified through traits that drive each species response to the environment (species functional traits, McGill et al., 2006; Westoby & Wright, 2006).
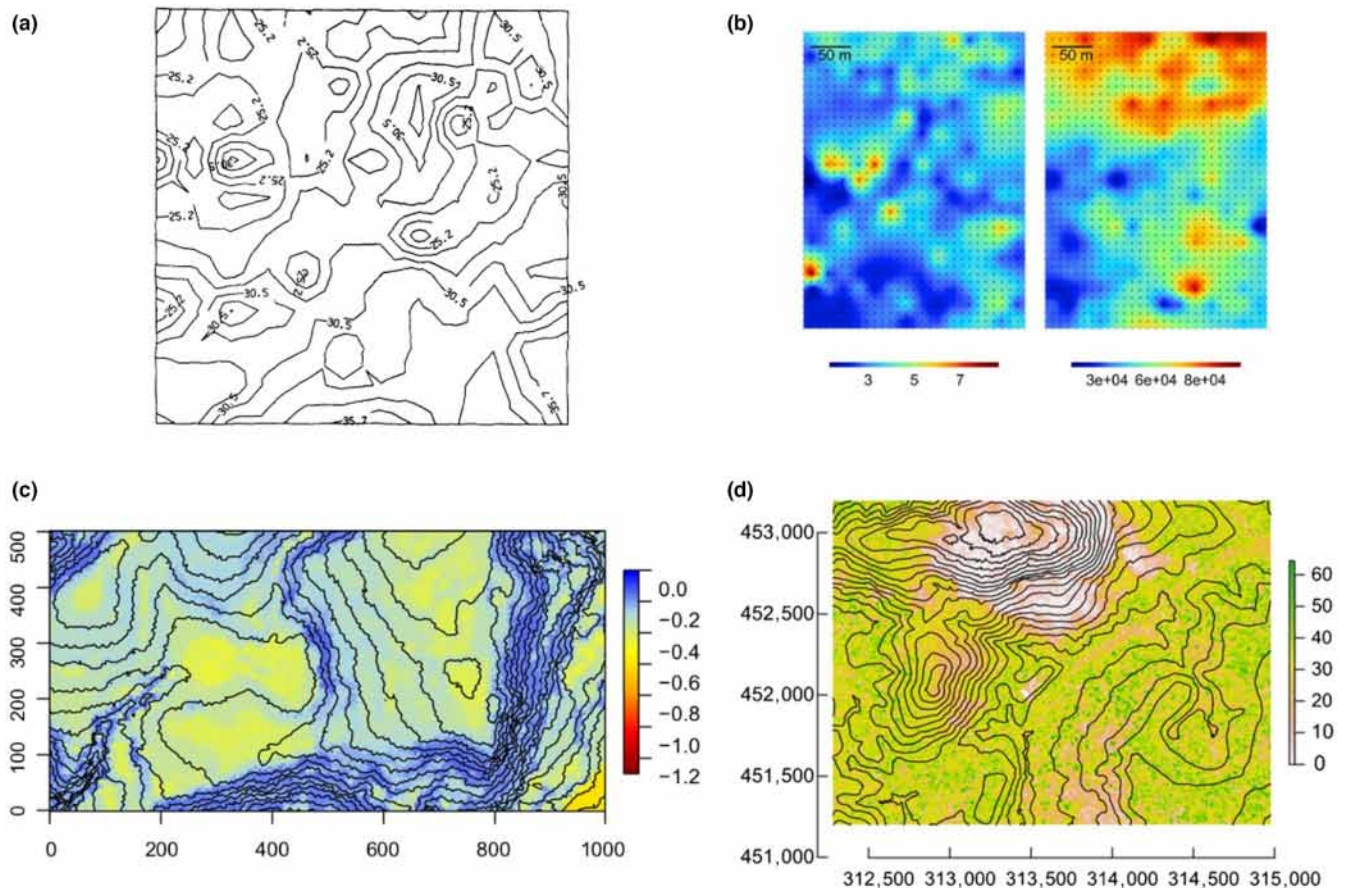
FIGURE 1 High environmental variability at a small spatial scale. (a) Soil nitrogen content in a 12×12 m plot at Cedar Creek (USA) in g kg^{-1}, Tilman, 1982; (b) Carbon in % (left) and aluminum in ppm (right) soil content in a 12-ha (250×500 m) plot at The Nouragues (French Guiana), Zinger et al., 2019; (c) Soil water content during mid-dry season of a regular year in MPa in a 50-ha (1000×500 m) forest plot at Barro Colorado Island (Panama), Kupers et al., 2019. Coordinates in m; (d) Canopy height in m and topography (10 m spaced elevation lines) in a 50-ha (2500×2000 m) area at the Nouragues Research Field Station, Tymen et al., 2017. Coordinates in m (UTM 22N).

Similar to the environment that presents highly dimensional variation at local scales, these functional species differences spread along many dimensions within communities (Baraloto et al., 2010; Hutchinson, 1957, 1959; Kraft et al., 2015; Maréchaux et al., 2020; Rüger et al., 2018; Vleminckx et al., 2021).

In this paper, we explore the potential that the role of environmental variation in shaping *observed* IV has been largely underexplored with important consequences on our understanding of the effect of *observed* IV on community dynamics. Indeed, a great part of *observed* IV might emerge from species responses to a high-dimensional environment (Figure 3): observed differences among individuals of the same species can be caused by the often poorly quantified response to the micro-environment they experience. This imperfect description can notably result from an insufficient set of environmental variables being measured, from a mismatch between the spatio-temporal scale of measurements and the one of individual attribute variation, or from measurement errors in both environmental variables and individual attributes. Apart from the latter, which can be unstructured and accounted for through specific tools (e.g., Calder et al., 2003), and that we do not explore here, the resulting variation among conspecific individuals would be structured in

space and time, and not necessarily caused by genetic variation. The study is divided into three parts. In the first part, we present a virtual experiment to illustrate the fact that large observed IV can emerge from environmental variation, having thus consequences on the structure of IV. In the second and third parts, we present insights from experimental clonal *Eucalyptus* plantations and tropical forest inventories in order to test if these results are supported by experimental and empirical data. For each of these three parts, we detail the corresponding material and methods as well as the results. We then provide a general discussion. Throughout the study, we examine three main hypotheses (Figure 4): (i) the large IV observed in natural communities can emerge from heterogeneity in multiple unobserved environmental dimensions which is often mischaracterized; (ii) because environmental variation is structured in space and time, IV is likely to be similarly structured as well, suggesting that it is not appropriate to represent IV as a purely random noise in models; and (iii) since a large observed IV does not necessarily imply that conspecific individuals substantially differ in their fundamental niche, conspecific individuals may still respond more similarly to environment than heterospecific individuals. We, therefore, call for a reconsideration of the nature and structure of IV, which could shed new
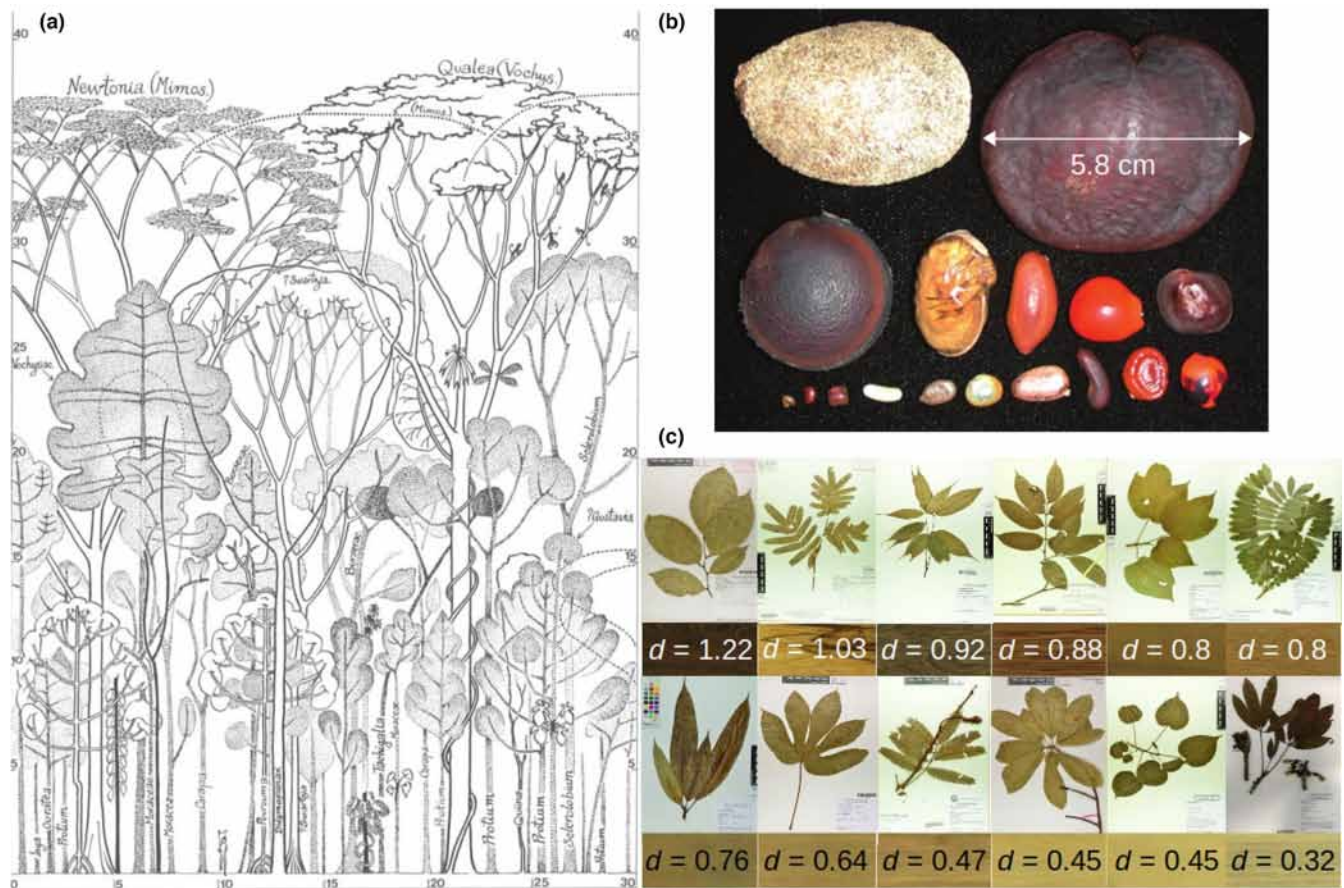
FIGURE 2 Morphological diversity of tree species illustrating strong differences between species. (a) Diversity of tree species architecture and height in a tropical forest (Hallé et al., 1978). Coordinates are in m; (b) Diversity of seed size and shape from 17 tree species of the Fabaceae family in the Peruvian Amazon (Muller-Landau, 2003); (c) Diversity of leaf size and shape (herbarium of Cayenne, Gonzalez et al., 2021) and of wood aspect (reflecting wood characteristics) and density (Normand et al., 2017) for 12 tree species in French Guiana. Species from top left to bottom right are *Bocoa prouacensis*, *Zygia racemosa*, *Vouacapoua americana*, *Eperua falcata*, *Bagassa guianensis*, *Hymenolobium excelsum*, *Mangifera indica*, *Sterculia pruriens*, *Parkia nitida*, *Couroupita guianensis*, *Hura crepitans*, and *Ceiba pentandra*. Black bars next to herbarium samples indicate the scale (10 cm).

light on the coexistence conundrum. While we acknowledge the existence of genetically based individual variations, and that plasticity has a genetic basis (Nicotra et al., 2010; Westerband et al., 2021), we suggest that a substantial part of *observed* IV might result from the higher dimensionality of the species niche and the finer scale of environmental variation than typically observed. Species differences along these many dimensions can lead to multiple local inversions of species hierarchy in an environment varying in space and time, thereby allowing the stable coexistence of numerous species.

2 | THEORETICAL ILLUSTRATION: UNOBSERVED ENVIRONMENTAL DIMENSIONS RESULT IN LARGE OBSERVED IV

We first conducted a virtual experiment to illustrate the hypotheses explored in this study. To do so, we generated simulated data of an individual attribute (e.g., tree growth) depending on a certain number of environmental variables varying in space, and then analyzed the simulated data assuming that most of the environmental variables are actually unobserved, as is typically the case in the field. This simulation experiment does not aim to accurately represent ecological reality, but to simply illustrate our points.

2.1 | A "perfect knowledge" simulation model

We considered a set of J species with I individuals each, distributed in a virtual landscape. The environment was assumed to be fully known and defined by N environmental variables, X_1 to X_N, that were each randomly and independently generated in the landscape, assuming spatial autocorrelation. These variables could also represent the independent axes that would result from, for example, a PCA of many environmental variables. Individual location was drawn randomly in a virtual landscape defined by a $C \times C$ square grid, each cell corresponding to a particular environment (Figure 5a). Individuals were identical within species (same model
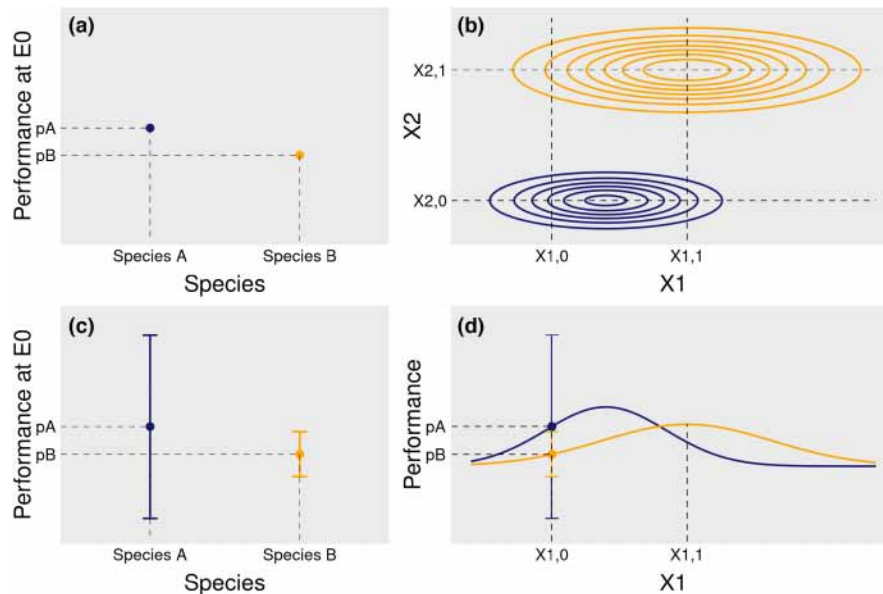
FIGURE 3 Reinterpreting observed intraspecific variability (IV): from niche widening to niche projection into a high-dimensional environment. In (a), within a given environment E0 defined along an environmental axis X1 (E0 = E(X1,0)), conspecific individuals are identical and have the same performance pA and pB, for species A (blue) and species B (orange). Species A outcompetes species B in E0. Actual measured differences among conspecific individuals, shown in (c), can be interpreted in different ways. First, as conspecific individuals exhibit contrasting attributes in E0, they become more different. This can result in some heterospecific individuals having similar performances: IV would make species less different. Alternatively, IV measured in E0 results from the variation of unobserved environmental variables (E0 = E(X1,0, X2); (b)). Contrasting performances among conspecific individuals in E0 do not result from intrinsic differences among them but from differences in the local environment they experience and that was poorly characterized, that is, the number of observed dimensions is lower than the actual number of environmental dimensions. Similarly, although species niches present some overlap when projected on one dimension (d), they do not overlap in the two-dimensional space (b). Moreover, while species A outcompetes species B on average when X1 = X1,0, the opposite occurs when X1 = X1,1 (d), leading to an inversion of species hierarchy between different environments. Similarly, while species A outcompetes species B in E(X1,0, X2,0), the opposite occurs in E(X1,0, X2,1). Although only two dimensions are shown, species respond to many environmental variables varying in space and time, multiplying the possibilities of niche segregation and hierarchy inversions between species, offering room for species coexistence in a variable high-dimensional environment. The code used to generate this figure is available online.

parameters for all conspecific individuals), but different between species (different model parameters between heterospecific individuals).

We considered the following "perfect knowledge" mathematical model, which depicts the exact attribute Y_{ijt} (e.g., growth) of an individual i of species j given its environment at time t (Equation 1, Appendix S1).

$$\ln\left(Y_{ijt}\right) = \beta_{0,j} + \beta_{1,j}\ln\left(X_{1,ijt}\right) + \beta_{2,j}X_{2,ijt} + \ldots + \beta_{N,j}X_{N,ijt} \quad (1)$$

In this model, $\beta_j = [\beta_{0,j}, \ldots, \beta_{N,j}]$ is the vector of parameters defining the response of individuals of species j to the environment. Because conspecific individuals respond similarly to environmental variables, variation in Y_{ijt} among them is only due to differences in the environment where and when each individual is growing. Using this model, we computed the attribute Y of the $I \times J$ individuals at T dates, assuming that values for some of the environmental variables changed between dates, and thus obtained a simulated dataset $\{Y_{ijt}, X_{1,ijt}, \ldots, X_{N,ijt}\}$ with $N = 10, I = 300, J = 2, C = 500$ and $T = 2$.

2.2 | An "imperfect knowledge" statistical model

Second, we considered an "imperfect knowledge" statistical model for which we assumed that only one explanatory variable X_1 (e.g., light) in the above simulated dataset has been measured at the relevant scale among all the environmental drivers that actually determine response variable Y (Equation 2, Appendix S1). This model represents the ecologist's imperfect understanding of attribute Y, as few sites would actually offer quantification of variation of several environmental variables at the individual scale that is relevant. The model includes a species fixed effects on the intercept and on the slope ($\beta'_{0,j}$ and $\beta'_{1,j}$) and a random individual effect $b_{0,i}$ on the intercept with $b_{0,i} \sim N(0, V_{bj})$, where V_{bj} is the intraspecific variance for species j. We estimated the model parameters based on the simulated dataset introduced above but considering only the first explanatory variable $\{Y_{ijt}, X_{1,ijt}\}$, the remaining "unknown" environmental effects being contained in the model residuals, ε_{ijt}.

$$\ln\left(Y_{ijt}\right) = \left[\beta'_{0,j} + b_{0,i}\right] + \beta'_{1,j}\ln\left(X_{1,ijt}\right) + \varepsilon_{ijt}, \varepsilon_{ijt} \sim N(0, V_j) \quad (2)$$

A combination of approaches

THEORETICAL
Model with clones in a multidimensional environment

EXPERIMENTAL
Clonal *Eucalyptus* plantations

EMPIRICAL
Tree inventories in three natural tropical forests

LITERATURE
Micro-environment, species differences

Insights on IV

Large IV can emerge from high dimensional environmental heterogeneity

IV does not imply that individuals are different, nor that species niches overlap

IV is structured in space and time

Despite IV, conspecifics answer more similarly than heterospecifics

Consequences

Species differ significantly in a high number of dimensions

Shift in species hierarchy in space and time
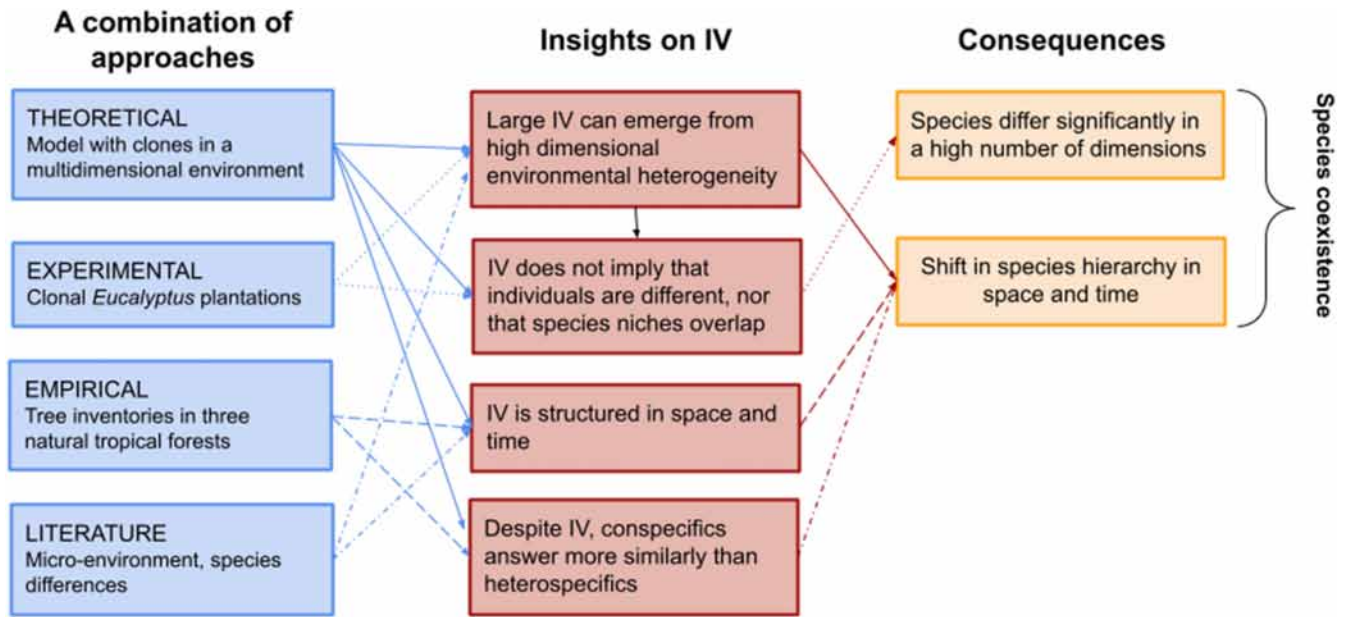
Species coexistence

FIGURE 4 Multiple insights on the nature of IV and its consequences on individual and species differences. We used literature and data analyses of various nature to support the hypothesis that a large part of observed IV can result from multidimensional environmental variations that are spatially and temporally structured rather than by intrinsic and spatio-temporally unstructured differences between conspecific individuals, with radically different consequences on species coexistence.

We used default priors for variance parameters (Student's T with three degrees of freedom, location = 0 and scale = 2.5), and normal distributions (with mean = 0 and variance = 1) for mean parameters. The estimation of model parameters was done using a Bayesian approach using Stan software with the brms R package (Bürkner, 2017, 2018). We used four MCMC chains with different initial values. We made 10,000 iterations for each chain with a warm-up period of 5000 steps and a thinning rate of one fifth. We obtained 1000 estimations per chain per parameter and examined trace plots, posterior distributions, R-hats, and ESS statistics to check convergence of the chains (Appendix S1).

2.3 | Apparent niche overlap and observed intraspecific variability as a result of unobserved environmental variables

Despite the fact that conspecific individuals were identical and species responses to environment were different, the variance estimates $\hat{V}_{bj}$ for individual random effects of species j were large, and species responses to the environment overlapped (Figure 5b). This is due to the contribution of the unmeasured variables $\{X_{2,ijt}, ..., X_{N,ijt}\}$ in determining the variation of Y across individuals.

Since it is driven by spatially autocorrelated variables (Equation 1), the response Y was spatially autocorrelated across conspecific individuals (Figure 6). This means that two neighboring conspecific individuals have a more similar attribute Y than two distant conspecific individuals. Additionally, the variance of Y was lower within than between species: conspecific individuals responded more similarly to the environment than heterospecific individuals did (Figure 6).

With this simulation experiment, we simply illustrated that: (i) apparent high IV can emerge due entirely to an imperfect description of the environment exclusively (e.g., unobserved environmental dimensions), (ii) the spatial structure of this IV follows the spatial structure of the underlying environmental variables, and (iii) this IV does not make species less different in their response to environmental variables (Figure 6) despite apparent niche overlap in one dimension (Figure 5b).

3 | EXPERIMENTAL INSIGHTS: LARGE OBSERVED INTRASPECIFIC VARIABILITY IN A CLONAL TREE PLANTATION

We then moved from a theoretical illustration to an experimental approach using census data from clonal *Eucalyptus* plantations, where genetic variability among individuals growing within a single same site is controlled. We explored the partitioning of IV between intrinsic (genotypes) and extrinsic sources, which is often infeasible in natural settings, to demonstrate that substantial observed IV can indeed emerge from genetically identical individuals in the field, even when persisting in an apparently homogeneous environment.

3.1 | An extreme case of controlled genetic and environmental variation

The EUCFLUX experiment (São Paulo state, Brazil) is a clonal trial with a replicated, statistically sound design (le Maire et al., 2019). It includes 14 genotypes of five different *Eucalyptus* species or hybrids
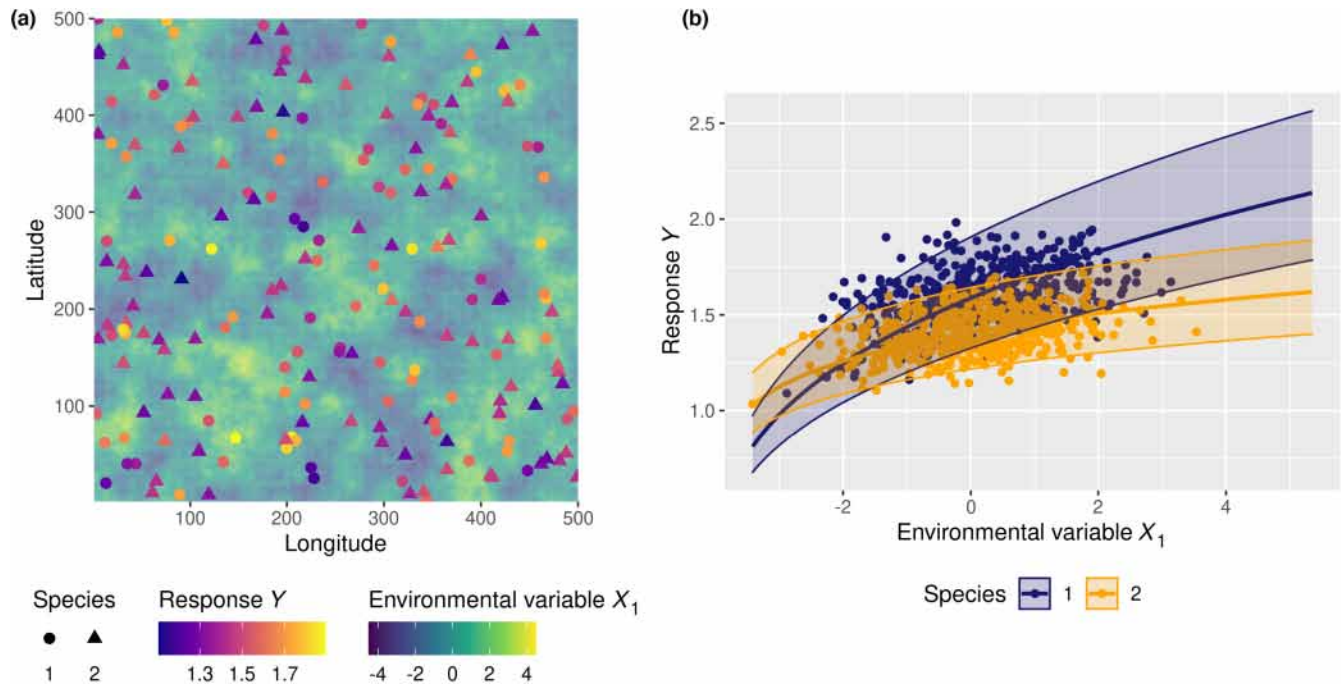
FIGURE 5 Observed intraspecific variability as a result of the imperfect characterization of the environment. A simulated response variable (Y, e.g., growth) is generated for individual clones of two species thriving in a high-dimensional environment. This response variable was first computed as a function of 10 environmental variables ("perfect knowledge" model, Equation 1), but is then analyzed using a statistical model that accounts for the unique environmental variable that was assumed to be observed in the field (X_1, e.g., light) and includes a random individual effect ("imperfect knowledge" model, Equation 2). The intraspecific variability estimated with these random individual effects is then due to the variation in space and time of the nine unobserved environmental variables. (a) Positions of a sample of $I = 600$ individuals from $J = 2$ species in a landscape defined by a square grid of $C \times C$ cells ($C = 500$). The background color indicates the value of the observed environmental variable X_1 on each cell at date t. The response Y of each individual, which depends on the environment within each cell (Equation 1), is also indicated by a color scale. (b) Response Y as a function of the observed environmental variable X_1 for the two species. Points represent the data $\{Y_{ijt}, X_{1,ijt}\}$. Thick lines represent the predictive posterior means for the two species. The envelopes delimited by two thin lines represent the 95% credible intervals of the predictive posterior marginalized over individuals (taking into account $\hat{V}_{bj}$). The envelopes thus represent the intraspecific variability which is due to the $N-1$ unobserved environmental variables.

of various origins. Each genotype is planted in plots of 100 trees, at a density of 1666 trees per hectare, and replicated spatially in 10 blocks (Figure 7). The experimental set-up was designed to minimize the variation in environmental factors among blocks, which were separated by less than 1.5 km within a homogeneous 200-ha stand showing small variation in soil properties. Tree diameter at breast height (D) has been measured over five complete censuses, spanning 6 years, age at which such plantation is generally harvested (see le Maire et al., 2019 and Appendix S2 for further details on this experimental set-up).

3.2 | A partitioning of observed variance among individual tree growth

We computed annual diameter growth (G) in mm·year^{-1} for each tree as well as a competition index (C) as the sum of the basal area of the eight direct neighbors of each tree. The dataset included 64,125 growth estimates corresponding to 13,531 trees in total. To quantify the relative importance of the different sources of growth variability,

we used a statistical hierarchical growth model (Equation 3) including an intercept (β_0), fixed effects of the log-transformed diameter (β_1) and competition index (β_2), and random effects on the intercept for the block ($b_{0,b}$, with $b_{0,b} \sim N(0, V_b)$), the genotype ($b_{0,g}$, with $b_{0,g} \sim N(0, V_g)$), the census date ($b_{0,t}$, with $b_{0,t} \sim N(0, V_t)$), and the individual ($b_{0,i}$, with $b_{0,i} \sim N(0, V_i)$). All the data were log-transformed and scaled, and a constant of 1 mm was added to all growth values to avoid undefined logarithms.

$$\ln(G+1)_{it} = [\beta_0 + b_{0,b} + b_{0,g} + b_{0,t} + b_{0,i}] + \beta_1 \ln(D)_{it} + \beta_2 \ln(C)_{it} + \varepsilon_{it}, \varepsilon_{it} \sim N(0, V) \quad (3)$$

We used default Student's T priors for variance parameters, and normal distributions (with mean = 0 and variance = 1) for mean parameters. Model parameters were estimated using the same Bayesian approach as for the statistical model of the theoretical illustration (Appendix S2).

We then examined the proportion of the model residual variance (variation of the response variable that is not explained by the covariates) related to each random effect in order to partition the block, genotype, date, and individual variances.
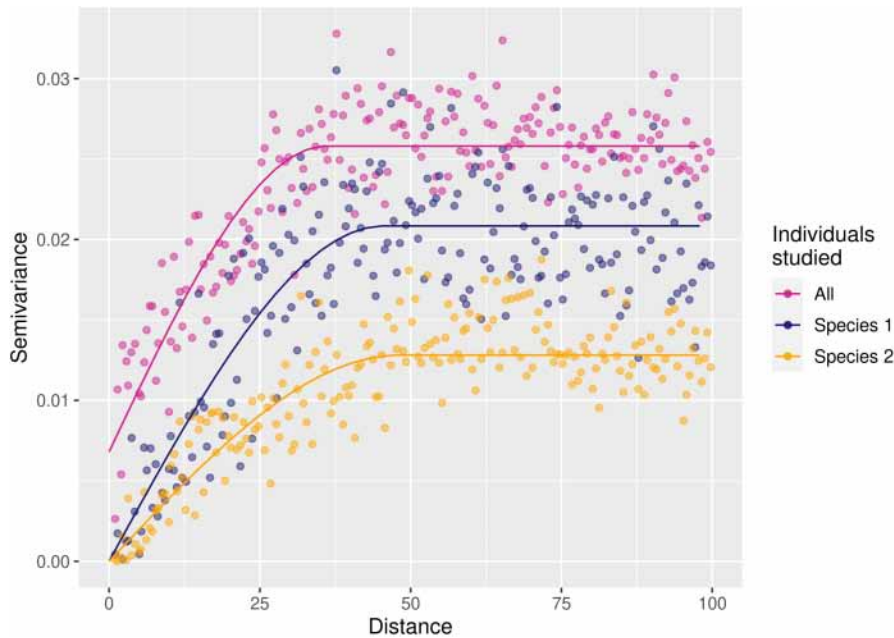
FIGURE 6 Spatial autocorrelation of attribute Y across individuals within and between species ($J = 2$) in a simulation experiment. This semivariogram represents the semivariance of the individual mean attribute Y as a function of the distance between individuals. The increasing curves evidence spatial autocorrelation in Y (similar results using Moran's I test). The semivariance of all individuals taken together (purple curve) is higher than the semivariance of conspecific individuals for the two species (orange and blue curves), which means that intraspecific variability is lower than interspecific variability.
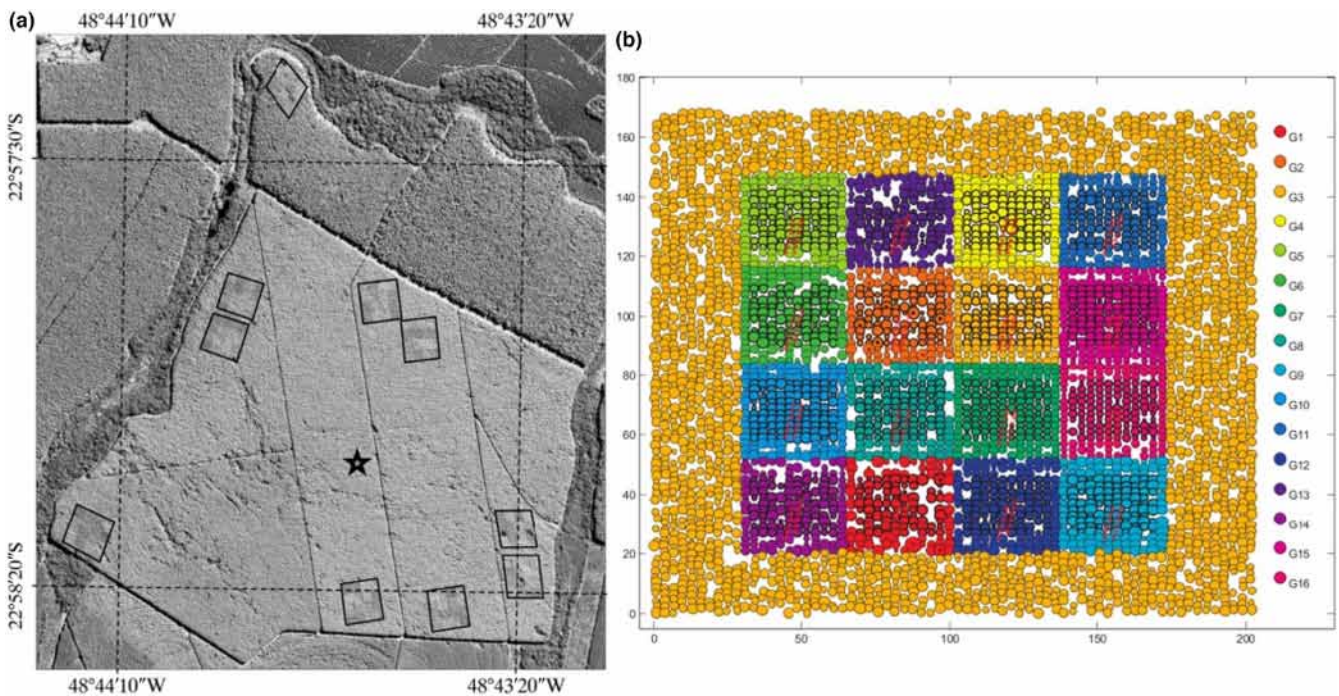
FIGURE 7 Experimental setup of the EUCFLUX experiment. The 10 blocks (a) and the organization of the 16 genotypes within a block (b). In our analyses, two genotypes were discarded because they were obtained from seeds and not clones and therefore included some genetic variability. A more complete figure legend can be found in le Maire et al., 2019.

3.3 | Variation among individuals is not explained by genotype

While minor variability was associated with blocks (Table 1), confirming that they are broadly homogeneous by design, the variability associated with temporal factors was predominant. This reveals that the competition index (C) used in the analysis to encapsulate the effect of progressive canopy closure does not fully encompass all temporal effects.

Importantly, the variability between individuals was almost twice as high as the variability between genotypes (Table 1). Hence, even in such an extremely conservative case, where environmental variation in space is minimized and genotypic variability controlled, a large part of measured IV cannot be explained by purely genetic differences among individuals that would remain independent of the environment as an IV simulated through independent random draws would be. This is in broad agreement with

TABLE 1 Mean posteriors of the *Eucalyptus* model and their estimation errors and residual variance partitioning among the different random effects.

	Intercept (β_0)	Diameter (β_1)	Competition (β_2)	Individual variance (V_i)	Block variance (V_b)	Genetic variance (V_g)	Temporal variance (V_t)	Residual variance (V)
Estimate	−3.50E-02	5.50E-01	−2.70E-01	2.30E-01	6.00E-02	1.30E-01	1.30E+00	5.10E-01
Estimation error	5.00E-01	5.00E-03	8.90E-03	4.00E-03	1.80E-02	3.10E-02	5.40E-01	2.00E-03
% unexplained variance				10.31%	2.69%	5.83%	58.30%	22.87%

common garden studies in community genetics, which typically find a trait heritability lower than 1 for genetically identical plants growing in a relatively homogeneous environment (e.g., Shalizi & Isik, 2019).

These results from a controlled experiment suggest an underestimated role of environmental micro-heterogeneity in shaping variation among individuals. This could be due to both unobserved environmental variables and a larger scale of environmental measurements (the plot) than the measured response (the individual scale). For instance inevitable spatial variation of biotic and abiotic variables (soil microbiome, pathogens, soil structure and water content, light, neighborhood interactions etc.) at fine scales (e.g., cm to m scale, hence impacting tree-scale environment, Baraloto & Couteron, 2010, Figure 1) as well as potential early manipulations of the young plant, the way it was planted, *etc*, could drive individual growth response.

4 | EMPIRICAL INSIGHTS: OBSERVED INTRASPECIFIC VARIABILITY IS HIGH AND SPATIALLY STRUCTURED AND DOES NOT "BLUR SPECIES DIFFERENCES" IN TROPICAL FORESTS

To explore some of our hypotheses in natural communities, we then used data from three long-term tree inventories in tropical forests, from Amazonia (Paracou, French Guiana; Gourlet-Fleury et al., 2004), Central America (Barro Colorado Island, Panama; Losos & Leigh, 2004) and South-East Asia (Uppangala, India; Pélissier et al., 2011). More specifically, we inferred observed IV, tested if individual growth showed local spatial autocorrelation, that is, was structured in space, and if conspecific individual growth was more similar than heterospecific individual growth locally. These three sites encompass contrasting climatic conditions (rainfall ranging from 2600 in BCI to 5100 mm·year^{-1} in Uppangala), disturbance regimes (incl. various logging experiments in Paracou), and topography (from gentle in BCI to mountainous in Uppangala), making them representative of the global tropical forests. The data from the tropical forest inventories that we used in this paper are summarized in Table 2.

For all three datasets, annualized growth between two censuses was computed as the difference of DBH (≥10 cm) between two consecutive censuses, divided by the time period between those two

censuses. Growth estimates < −2 or > 100 mm·year^{-1} as well as individuals from incompletely identified species and individuals and species with a single observation were discarded prior to analysis. Mean annual growth for each individual tree was then computed as the difference of DBH between the first and the last time a tree was measured, divided by the time period between those two measurements.

4.1 | High observed intraspecific variability in tree growth in tropical forests

To quantify the relative importance of intra- versus interspecific variability in each site, we used a hierarchical growth model (Equation 4), including an intercept β_0, a diameter (D) fixed effect β_1, a species random effect $b_{0,j}$ (with $b_{0,j} \sim N(0, V_b)$) and an individual random effect $b_{0,i}$ (with $b_{0,i} \sim N(0, V_{bi})$) on the intercept. All data were log-transformed and scaled, and a constant of 2 mm was added to all growth values to avoid undefined logarithms.

$$\ln\left(G_{ijt} + 2\right) = \left[\beta_0 + b_{0,j} + b_{0,i}\right] + \beta_1 \times \ln\left(D_{ijt}\right) + \varepsilon_{ijt}, \varepsilon_{ijt} \sim N(0, V) \quad (4)$$

We used default Student's T priors for variance parameters, and normal distributions (with mean = 0 and variance = 1) for mean parameters. We estimated the inter- and intraspecific growth variability from the variance of the species (V_{bj}) and individual (V_{bi}) random effects, respectively. Model parameters were estimated using the same Bayesian approach as before (Appendix S3).

For the three sites, IV estimated from the growth model (V_{bi}, ranging from 0.41 to 0.66) was of the same order of magnitude as the interspecific variance (V_{bj}, ranging from 0.37 to 0.67) (Table 3). Overall, a large share of the variability in tree growth comes from individual effects in the three sites, even after accounting for the effect of diameter on tree growth, showing a high intraspecific variability in growth in these tropical forests.

4.2 | Spatial autocorrelation of individual growth within species at the local scale in tropical forests

To test whether individual growth was spatially autocorrelated, we performed in each site, spatial analyses of the mean individual growth values. We chose a conservative approach based on mean

TABLE 2 Features of the three tropical forest datasets used as empirical case studies.

Site	Rainfall (mm·year⁻¹)	Sampling	Min DBH	Nb of censuses	Periodicity	Disturbance	Topography	Nb of species	Nb of individuals	Data source
Paracou, French Guiana	3000	15×6.25 ha (incl. 12 logged plots)	10 cm	24	1-2 years since 1992	Natural disturbances + selective logging	Flat	613	69,548	Hérault and Piponiot (2018)
BCI, Panama	2600	50 ha	10 cm	8	5-y since 1980	Natural disturbances	Hilly	225	37,224	Condit et al. (2019)
Uppangala, India	5100	5,92 ha (4 transects and 3 plots)	9.5 cm	20	1-year since 1992	Natural disturbances	Mountainous	102	3789	Le Bec et al. (2015)

individual growth without accounting for the effect of diameter, thus without removing ontogenetic differences and considering the pattern of individual growth as it is in the field. More specifically, we performed Moran's I one-tailed tests as implemented in the *ape* R package (Paradis & Schliep, 2019), for pairs of conspecific individuals less than 100 m apart in the same plot (to avoid capturing the effect of treatment in Paracou and including the spaces between the plots). For the most abundant species, we sampled 3000 individuals with a uniform probability. We considered only the species with more than five conspecific neighbors <100 m apart in the same plot.

Positive spatial autocorrelation in tree growth between conspecific individuals was significant for 19%–31% of the species in the three sites, representing between 45 and 79% of the total number of individuals (Table 4). Spatial autocorrelation was however much higher in logged plots as compared to unlogged in Paracou, because of a more heterogeneous light environment resulting from logging history (Appendix S3). Note that the absence of significant spatial autocorrelation for some species is partly explained by their low abundance (see Figure S3.11 in Appendix S3).

4.3 | Higher similarity of growth within conspecific- than heterospecific individuals locally in tropical forests

To test if the performance of conspecific individuals was locally more similar than the performance of heterospecific individuals in the three sites, we also used mean individual growth, thus ignoring ontogenetic differences. We computed the mean individual growth semivariance (Baraloto & Couteron, 2010) considering either conspecific or heterospecific neighbors within a 100-m radius. In the first case, semivariance was estimated as the mean of the squared difference in individual mean growth for all pairs of conspecific individuals. In the second case, semivariance was estimated as the mean of the squared difference in individual mean growth for all pairs of individuals with an individual of the focal species and one of another species. We considered only the species with more than five individuals, and with more than five heterospecific neighbors within the 100-m neighborhood distance. For each species, we then compared the semivariances between conspecific and heterospecific individuals using a Mann–Whitney test with a 0.05 alpha-risk.

The mean individual growth semivariance appeared significantly higher among heterospecifics than among conspecific individuals for 42%–61% of the species in the three sites, representing 58%–89% of the total number of individuals (Table 5, see also Figure S3.11 in Appendix S3 regarding species with non-significant test). To control for a potential effect of species abundance on the semivariance estimations, we replicated the analysis by sampling a maximum of 10 individuals per species. The results were qualitatively unchanged (Appendix S3).

With these results obtained with empirical data, we showed that (i) observed intraspecific variability can be high in tropical forests, (ii) individual growth can be spatially structured, and (iii) conspecific

TABLE 3 Mean posteriors of the tropical forest model and their estimation errors and residual variance partitioning among the different random effects.

	Intercept (β_0)	Diameter (β_1)	Species variance (V_{bj})	Individual variance (V_{bi})	Residual variance (V)
Paracou					
Estimate	2.30E-03	−2.30E-01	5.20E-01	5.50E-01	7.50E-01
Estimation error	2.30E-02	2.80E-03	1.70E-02	2.20E-03	5.70E-04
% Variance			28.57%	30.22%	41.21%
Uppangala					
Estimate	8.20E-02	1.90E-01	3.70E-01	6.60E-01	5.90E-01
Estimation error	4.70E-02	1.30E-02	4.40E-02	8.60E-03	1.90E-03
% Variance			22.84%	40.74%	36.42%
BCI					
Estimate	1.90E-01	-2.20E-02	6.70E-01	4.10E-01	8.10E-01
Estimation error	4.70E-02	4.50E-03	3.50E-02	4.10E-03	2.00E-03
% Variance			35.45%	21.69%	42.86%

TABLE 4 Spatial autocorrelation of the growth of conspecific individuals in three tropical forest sites.

	Significant	Not significant
Paracou		
% Species	31.00	69.00
% Individuals	78.90	21.10
Uppangala		
% Species	18.50	81.50
% Individuals	45.30	54.70
BCI		
% Species	20.10	79.90
% Individuals	54.70	45.30

Note: Shown are the proportion of species, and of corresponding individuals, in percent, for which individual growth among conspecific individuals is significantly positively spatially autocorrelated. The spatial autocorrelation of individual growth was tested using Moran's I index.

individuals can have a more similar growth than heterospecific individuals at local scales. Altogether this suggests the signature of environmental variation that was not accounted for but that typically influences individuals at fine scale in many ways.

5 | DISCUSSION

5.1 | High-dimensional environmental variation leads to large observed intraspecific variability

IV can result from intrinsic differences among individuals or from extrinsic environmental variation, including biotic factors, or interactions of both (Moran et al., 2016; Violle et al., 2012; Westerband et al., 2021). While much emphasis has been placed on genetically driven IV in studies on coexistence (Barabás & D'Andrea, 2016; Booth & Grime, 2003; Ehlers et al., 2016), sometimes implicitly

through the use of independent random draws across individuals (Crawford et al., 2019; Hart et al., 2016; Lichstein et al., 2007), and although we acknowledge its ecological and evolutionary importance, we here argue that the importance of environmentally driven IV in natural communities has been underestimated and has radically different consequences for species differences and community assembly. More specifically, we argue that a large part of observed IV can result from an imperfect description of environmental variation in space and time.

First, using a simple simulation experiment, we illustrated how environmental variation in unobserved dimensions of the environment can produce large observed IV, although conspecific individuals are clones (Figure 5). Similarly, the variance partitioning of individual tree growth within a common garden of _Eucalyptus_ clones (le Maire et al., 2019) shows that the variance in growth between individuals is about twice as high as the variance between genotypes (Table 1). This reveals that a large part of the observed IV can emerge from nongenetic determinants, for example, micro-environmental variations that are poorly characterized, even when the variation of the environment was sought to be minimized.

Importantly, because IV can emerge from environmental heterogeneity and without underlying genetic differences, observed IV does not necessarily imply that conspecific individuals substantially differ in their response to the environment, nor that species niches overlap (Figure 3). Instead, large observed IV can also reflect the projection of species' high-dimensional niches within a high-dimensional environment that is variable in time and space: conspecific individuals differ with each other because they each thrive in a different microenvironment. In empirically observed data, such an IV is, therefore, the result of projecting a high-dimensional response (e.g., physiological processes), which is controlled by multiple macro- and microenvironmental variables, down to a low-dimensional, integrative response (e.g., annual growth) that is poorly characterized because of an incomplete view of the environmental variables that contribute to it. This reassigns an important part of observed variation among individuals, often represented as neutral or random

	Intraspecific variability < interspecific variability (i)	Intraspecific variability ~ interspecific variability (ii)	Intraspecific variability > interspecific variability (iii)
Paracou			
% Species	60.70	40.70	0.67
% Individuals	88.80	10.90	0.28
Uppangala			
% Species	42.20	62.20	4.44
% Individuals	57.70	23.60	18.80
BCI			
% Species	46.10	47.80	3.14
% Individuals	76.00	19.30	4.69

Note: The variability was estimated with the semivariance and the comparison was performed with a Mann–Whitney's test. The semivariances were computed for all species with >5 individuals and >5 heterospecific neighbors within 100 m in the same plot, and considering pairs of individuals that were less than 100 m apart and in the same plot. Shown are the proportion of species, and of corresponding individuals, for which (i) intraspecific variability was significantly lower than interspecific variability, (ii) intraspecific variability was significantly higher than interspecific variability, or (iii) the difference between inter- and intraspecific variabilities was not significant.

since they are seemingly unrelated to the observed dimensions of the environment (Figures 3 and 5, Table 1), to the classical niche theory (Hutchinson, 1957) and is in agreement with natural history observations of individual trait differences that are associated to species-specific ecological strategies.

The "biodiversity paradox" highlights that a large number of species can coexist while competing for a limited number of resources (Hutchinson, 1961). This puzzling question has generally been tackled considering trade-offs along a limited number of niche axes, often corresponding to resources (Rees, 2001; Tilman, 1982). But if the number of resources may indeed be relatively limited (e.g., light, water, and nutrients for plants), the number of independent environmental factors (e.g., microclimatic variables) that drive the performance of individuals for a particular level of resources is not. Environments are known to vary along multiple dimensions at fine scales in space and time (Figure 1), and in many cases, this variation has been shown to influence individual attributes (e.g., Fortunel et al., 2020).

Nevertheless, despite technological advances, many of these abiotic and biotic environmental factors are still poorly understood and monitored. As a result, the dimensionality of field observations is typically low compared with the high dimensionality of the environment in nature and the scale of environmental measurements is often coarser than the scale of the actual environmental variation (Bramer et al., 2018; Estes et al., 2018). The variability in individual attributes due to the variation of environmental variables that is not characterized therefore remains mostly a black box, and is typically summarized in terms of residual variance in statistical models (Albert et al., 2012; Siefert et al., 2015) or encapsulated into so-called "individual random effects" (Clark et al., 2007). We here emphasize that even in the absence of any intrinsic differences among conspecific individuals, a large IV can emerge from the imperfect characterization

of the environment (Figures 3 and 5, Table 1), which varies in a high number of dimensions at fine scales (Figure 1).

5.2 | Intraspecific variability is structured in space and time

IV has commonly been modeled through independent random draws around the species mean in community ecology studies (Barabás & D'Andrea, 2016; Courbaud et al., 2012; Hart et al., 2016; Lichstein et al., 2007; Uriarte & Menge, 2018). While this representation typically results from a lack of knowledge, with randomness being used as a substitute for more detailed understanding of underlying ecological processes (Clark et al., 2007), it encapsulates strong hypotheses relating to the nature of IV that are rarely discussed. In contrast, we emphasize here that IV is generally non-random and structured in both space and time.

At a given time, conspecific individuals that are distributed across space can strongly vary in their attributes (Moran et al., 2016; Poorter et al., 2018; Siefert et al., 2015; Violle et al., 2012). While this spatial IV has often been represented as random in community dynamics models, that is, implying that conspecific individuals can perform differently within the same environment (Figure 3c), a large part of this variability appeared in fact structured in space and likely associated with fine-scale spatial changes in the environment (Figure 3b, Moran et al., 2016). In our illustrative simulation experiment, the attribute of conspecific individuals varies spatially as a result of the environmental variation in space, and the spatial autocorrelation of conspecific attributes reflects the spatial autocorrelation of the environmental variables (Figures 5 and 6). While in our illustrative example, the attribute could be any trait or proxy related to performance, in our data

analyses, only growth was investigated. We hypothesize that the relationship between the investigated attribute and environmental variables drives the spatial autocorrelation due to environmental structure; thus so-called "response traits" (i.e., influenced by the environment, Lavorel & Garnier, 2002) should show similar spatial patterns than the ones here observed with growth. Moreover, the more an attribute integrates multiple processes (as for growth), the more likely it is to be controlled by the environment, and therefore should have a structured spatial pattern. We further hypothesize that the more related to performance an attribute is, the more likely it is to show spatial patterns due to local adaptation. Determining which and how traits are related to performance remains one of the main research challenges of functional ecology (Brodribb, 2017; Shipley et al., 2016; Violle et al., 2007).

Similarly, data from three long-term forest inventory sites across the tropics revealed spatial autocorrelation in tree diameter growth of conspecific individuals (Table 3), the insignificance of spatial autocorrelation for some species being constrained by their low abundance (Appendix S3). These results suggest that IV is strongly driven by the spatial variation of the environment, which is itself highly structured (Figure 1). However, we acknowledge that genetically driven IV can also be spatially structured, for instance via dispersal patterns or natural selection (Moran et al., 2016). We hypothesize that in that case, attributes would likely be randomly structured in space (Getzin et al., 2014) or correlated at the spatial scale of seed dispersal, typically several tens of meters in tropical forests (Clark et al., 2004; Muller-Landau et al., 2008; Seidler & Plotkin, 2006), while environmental variables are typically highly spatially correlated at fine scales (e.g., meter scale, Baraloto & Couteron, 2010). We also acknowledge that natural selection can happen at fine scales (Marrot et al., 2021), and could thus produce spatially structured IV due to local genetic adaptation. Nevertheless, data documenting genetic variation within species can still reveal higher similarity between conspecific than heterospecific individuals locally as well as nonoverlapping species niches (Schmitt et al., 2021). Importantly, any local genetic adaptation does not preclude that the imperfect characterization of environmental variations generates large observed IV that is structured in space and time and whose consequences cannot be well represented and understood using a random variation around a species mean.

In communities of sessile organisms such as trees, IV has been commonly structured in space using individual random effects, which vary among conspecific individuals but stay constant through the lifetime of individuals (Clark et al., 2007; Vieilledent et al., 2010). We here argue that while this approach can *reveal* the spatial structure of IV through inference, the use of the resulting estimated standard deviation term to introduce individual variation in simulations of community dynamics is not sufficient to *produce* a spatially structured IV, as we showed is observed in natural communities. Similarly, individual attributes can change over time. Because individuals within a species can be measured at different points in time, as it is often the case when assembling functional trait databases, for

example (Albert et al., 2011; Kattge et al., 2020; Zanne et al., 2009), this can lead to observed IV with no observable structure when it is only characterized by a variance around a species mean (Figure 3c). But a large part of this observed IV is actually structured in time and associated with temporal changes in the environment. For instance, the temporal storage effect (Chesson & Warner, 1981), a well-known coexistence mechanism, structures species performance because species are able to "store" growth during favorable timespans to overcome lean times; mast-seeding or masting, which describes periodic and synchronized massive seed production of conspecific individuals, would also result in a temporally structured IV (Koenig & Knops, 2005). Temporal variation in individual response within a species can typically be structured with temporal random effects (Clark et al., 2007). Temporal random effects have been used to estimate the inter-annual variability in tree growth (Fortunel et al., 2018; Metcalf et al., 2009) and fecundity (Clark et al., 2007) for example. In all those examples, temporal environmental variation affects conspecific attributes in the same way (Clark, 2010).

Overall, the representations of IV rarely reflect the overwhelming empirical observations that IV is spatio-temporally structured by the environment, ontogeny, local adaptation, and interactions between those factors. We therefore call for a reconsideration of the nature and the way of integrating IV into models of community dynamics. When IV is modeled randomly with a variance around a species mean, it implies that conspecific individuals can perform differently in the exact same environment, thus implying intrinsic differences between conspecific individuals. This type of unstructured IV can result in an overestimated increase in species niche overlap, which makes species more similar in their response to the environment (Figure 3a,c, Stump et al., 2021). While trait heritability has rarely been considered in studies on the role of IV on coexistence (but see Barabás & D'Andrea, 2016), in some studies, the random variation in attributes across conspecific individuals is considered as environmental, because it is not heritable in the model (e.g., Lichstein et al., 2007; Moran et al., 2016). However, environmentally driven IV should be structured in space and time, as the environment is (Figures 1 and 6). In addition, when IV is randomly distributed among conspecific individuals, similarity among conspecific individuals is systematically underestimated, which is not the case when IV is structured in space and time (Banitz, 2019; Purves & Vanderwel, 2014), as discussed hereafter.

5.3 | Conspecific individuals respond more similarly than heterospecific individuals locally

Species differ in multiple attributes, responding to a high number of environmental variables (Figure 2), but often in ways that cannot be readily observed. If observed IV results mainly from high-dimensional environmental variation in space and time rather than from intrinsic differences between conspecific individuals, then for a given environment, conspecific individuals should respond more similarly than heterospecific individuals. This is the case in our

illustrative simulation experiment, where the fact that conspecific individuals have exactly the same set of parameters and respond identically to spatial and temporal changes in the environment results in higher inter- than intraspecific variance in the response locally (Figures 5 and 6).

Corroborating this point of view, pairs of conspecific individuals in 11 North American temperate forest stands showed higher correlation in their temporal variation of growth rate or fecundity than pairs of heterospecific individuals on average (Clark, 2010). This indicates that conspecific individuals responded more similarly to environmental variation in time than individuals of different species. Importantly, these results were obtained in a system with high observed IV (leading to an apparent species niche overlap), where species responded in the same direction to environmental changes (e.g., increased tree growth in climatically favorable years). Hence, considering the temporal structure of IV revealed species differences that were not apparent otherwise, since they led to spreading along a high number of dimensions that varied at fine scales (Clark, 2010). However, as well highlighted by Stump et al. (2021), these results were often misinterpreted as an evidence that IV fostered coexistence. As another piece of evidence presented here, pairs of spatially proximal conspecific individuals tended to present more similar temporal means in absolute tree growth than pairs of close heterospecific individuals across three large contrasted tropical forest sites (Table 5). This provides new empirical evidence that, although estimated IV can be substantial, conspecific individuals respond more similarly than heterospecific individuals to environmental variation in space.

A stronger similarity in the response to environment between conspecific than heterospecific individuals locally is the signature of differences in species response to the environment. As a result environmental variation in space and time leads to local or punctual inversions of species hierarchy in performance (Figure 3d). As possibilities of hierarchy inversions between species increase rapidly with increasing dimensionality (Figure 3b) and variability of the environment, the high-dimensionality of the environment offers room for the stable coexistence of numerous species (Falster et al., 2017; Rüger et al., 2018). In the end, we therefore argue that a substantial part of IV is not a mechanism for coexistence in itself but can rather be the signature of species differences and environmental variation that allow coexistence: the high-dimensional species differences, which make them respond differently in a high-dimensional environment varying in space and time, can only be observed at the individual scale. In the absence of precise information on the many dimensions across which species differ and environment varies, large observed IV is the evidence of the niche mechanisms enabling species coexistence.

5.4 | Recommendations and concluding remarks

Most of the theoretical studies that have explored the role of IV in species coexistence so far did so by adding variances around species-specific means, thus considering IV as stochastic, which implies that conspecific individuals perform differently in the same environment. Here, we provide insights suggesting that large observed IV can emerge from environmental heterogeneity and is structured in space and time. We stress that this interpretation has strong consequences on the understanding of the effects of IV on species coexistence: (i) observed IV does not necessarily imply that conspecific individuals are strongly intrinsically different nor that species niches overlap, and (ii) the spatial and temporal structure of observed IV reveals stronger correlations of individual responses within species at local spatio-temporal scales, which reveals species niche differences in many dimensions. We thus call for a reconsideration of the nature of IV and of the way it is integrated in models, by thoroughly distinguishing its sources (intrinsic vs. extrinsic, and their interactions).

We acknowledge the existence of genetically driven IV, potentially due to local adaptation to the microenvironment, and its eco-evolutionary importance, but suggest that multidimensional environmental variation generates a large observed IV that is structured in space and time. We underline that environmentally driven structured IV has been largely overlooked in previous community ecology studies and has consequences on community dynamics which cannot be represented and understood using a random variation around a species mean. To this end, we recommend that empirical studies explore further the spatio-temporal structure of IV and how it relates to environmental variation along multiple dimensions, the differences in IV structure between species, and, when possible, assess the relative importance of genetically and environmentally driven IV, for instance by means of common garden experiments. Models of community dynamics should then endeavor to structure IV in space and time so that it reflects the high-dimensional variation in both the environment and species attributes, and not only some intrinsic differences between conspecific individuals (Banitz, 2019; Moran et al., 2016; Purves & Vanderwel, 2014). In both empirical studies and models, this implies that the species attributes are measured at the individual level, localized in space, and repeatedly observed in time. Adding spatio-temporal structure into the unexplained individual response in a community dynamics model will not properly substitute for the role of imperfectly characterized environmental variables however, particularly regarding the effect of environmental filtering. In particular, much care should be taken not to undermine the explanatory power of environmental variables in models when they are informed at a coarser scale than the individual scale, as it has been reported (Smart et al., 2021). Simultaneously, the monitoring of multiple environmental variables at fine scales in space and time is required in order to better capture their effect on individual attributes (such as physiological or mechanistic traits, Brodribb, 2017; Shipley et al., 2016), hence reducing the part of unexplained IV, and ultimately to better characterize the high-dimensionality of species niches. Altogether, these recommendations will enable to better account for species differences that are expressed at the individual level and evidence their impacts on the community dynamics in natura and in silico.

AUTHOR CONTRIBUTIONS

Camille Girard-Tercieux: Conceptualization (equal); data curation (equal); formal analysis (lead); writing – original draft (equal); writing – review and editing (equal). **Isabelle Maréchaux:** Conceptualization (equal); funding acquisition (equal); investigation (equal); methodology (equal); supervision (equal); validation (equal); writing – original draft (equal); writing – review and editing (equal). **Adam T. Clark:** Conceptualization (equal); investigation (equal); validation (equal); writing – review and editing (equal). **James S. Clark:** Conceptualization (equal); investigation (equal); validation (equal); writing – review and editing (equal). **Benoît Courbaud:** Conceptualization (equal); investigation (equal); validation (equal); writing – review and editing (equal). **Claire Fortunel:** Conceptualization (equal); investigation (equal); validation (equal); writing – review and editing (equal). **Joannès Guillemot:** Conceptualization (equal); investigation (equal); validation (equal); writing – review and editing (equal). **Georges Künstler:** Conceptualization (equal); investigation (equal); validation (equal); writing – review and editing (equal). **Guerric le Maire:** Conceptualization (equal); investigation (equal); validation (equal); writing – review and editing (equal). **Raphaël Pélissier:** Conceptualization (equal); investigation (equal); methodology (equal); supervision (equal); validation (equal); writing – review and editing (equal). **Nadja Rüger:** Conceptualization (equal); investigation (equal); validation (equal); writing – review and editing (equal). **Ghislain Vieilledent:** Conceptualization (equal); funding acquisition (equal); investigation (equal); methodology (equal); supervision (equal); validation (equal); writing – original draft (equal); writing – review and editing (equal).

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CONFLICT OF INTEREST STATEMENT

DATA AVAILABILITY STATEMENT

The appendices and all the code used for this study are available in a GitHub repository (https://github.com/camillegirardtercieux/co-exIV) and have been permanently archived on Zenodo (https://doi.org/10.5281/zenodo.5504013).

No new data were used in this study. For access to forest plot inventory data and *Eucalyptus* plantation data used in this study, refer to the data used in Le Bec et al. (2015), Hérault and Piponiot (2018), Condit et al. (2019), and le Maire et al. (2019). However, the analyses and reflections presented here are original.

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SUPPORTING INFORMATION

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