Opinion piece

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The use of fire to preserve biodiversity under novel fire regimes

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Novel fire regimes are emerging worldwide and pose substantial challenges to biodiversity conservation. Addressing these challenges and mitigating their impacts on biodiversity will require developing a wide range of fire management practices. In this paper, we leverage research across taxa, ecosystems and continents to highlight strategies for applying fire knowledge in biodiversity conservation. First, we define novel fire regimes and outline different fire management practices in contemporary landscapes from different parts of the world. Next, we

synthesize recent research on fire use and biodiversity, and provide a decision-making framework for biodiversity conservation under novel fire regimes. We recommend that fire management strategies for preserving biodiversity should consider both social and ecological factors, iterative learning informed by effective monitoring, and developing and testing new management actions. An integrated approach to learning about fire and biodiversity will help to navigate the complexities of novel fire regimes and preserve biodiversity in a rapidly changing world.

This article is part of the theme issue 'Novel fire regimes under climate changes and human influences: impacts, ecosystem responses and feedbacks'.

1. Introduction

Humans have long used fire to shape environments and meet societal needs [1,2]. For thousands of years, this has included manipulating the timing and location of ignitions and the amount and distribution of fuels [3]. More recently, human actions have also started to modify biotic distributions and climate patterns at global scales, influencing fuels and their availability to burn. These changes are already causing shifts in terrestrial ecosystems and contributing to biodiversity declines globally [4].

Fire regimes reflect patterns of recurrent fires and can be characterized by properties such as frequency, intensity, severity, spatial pattern and season [5,6]. Fire regime properties vary among ecosystems; for example, some tropical grasslands experience annual fires, while boreal forests might not experience fires for centuries. There is growing evidence that humans are altering the historical range and variation of fire patterns [7], and inducing environments different from those that humanity has experienced [8,9]. Deviations from historical fire patterns—which we refer to as novel fire regimes—can include increased or reduced fire activity [10,11]. While the causes and consequences of novel fire regimes are contentious, there is a consensus that multiple drivers are in play, including climate change, land use, biotic mixing and their underlying societal causes [12].

Understanding fire regimes and how they are changing is important because they modulate biodiversity across scales—from species to whole ecosystems [13,14]. In forests and woodlands with frequent high-intensity fires, plants have evolved fire-adaptive traits such as serotiny (retention of seeds on mature plants), smoke-induced germination and post-fire resprouting [15–17]. Fire regimes in these habitats can maintain open-canopy habitats, promoting diverse grass species favoured by sunlight [18] and, in turn, a variety of insects [19], land snails [20] and birds that thrive in fire-created niches [21]. Fire patterns can also contribute to maintaining a diversity of ecosystems [22]; for example, reduced fire frequency can transform open tropical savannas into closed forests [23,24].

Fire suppression in fire-prone areas has reduced fire frequency but led to fuel accumulation and more intense and severe wildfires [25]. Climate change is already amplifying fire danger, causing extensive wildfires in atypical seasons, elevations and vegetation types [26]. This rapid change is stressing species and ecosystems, even in fire-adapted regions [27]. Thus, understanding fire regime properties that preserve biodiversity under both natural and anthropogenic pressures is essential [28], while also prioritizing human health, safety and key ecosystem services like carbon storage or clean water supply [29].

Predicting how novel fire regimes influence biodiversity presents challenges [30]. Despite improvements in data collection and advances in modelling, the social and ecological complexities of fire-biodiversity relationships make it difficult to predict changes under new and unprecedented conditions [31]. Fire and biodiversity models are usually correlative, and rarely process-based, which complicates forecasting ecosystem dynamics and land use changes [32]. The stochastic nature of fire introduces further uncertainty. Integrative approaches, including testing the strategic use of prescribed burning, fuel management, wildfire suppression and other land management practices, require cutting-edge models and participatory approaches incorporating local knowledge and values.

This paper synthesizes scientific knowledge across taxa, ecosystems and continents to identify strategies for applying and managing fire in environments that differ those experienced by biodiversity in the past. It begins by examining the main approaches to fire use, their biodiversity impacts and the challenges and limitations of these practices. It then introduces a decision-making framework to guide fire management for biodiversity conservation. This integrative approach to decision-making emphasizes the importance of clear management objectives, understanding biodiversity outcomes of alternative management strategies and iterative learning to address the challenges posed by novel fire regimes.

2. Fire use and practices

(a) Broad approaches to fire use and management

One approach for maintaining biodiversity in fire-prone landscapes is to attempt to reinstate historical or reference fire regimes [33–36]. This can mean re-establishing fire as a natural ecological process. For instance, in North American longleaf pine ecosystems, a common objective is to revert to a pre-industrial fire regime, by considering as references the burning practices of Indigenous peoples and the role of natural ignitions [33,34]. Understanding the historic range and variation of fire regimes can broaden fire use possibilities and guide management towards more resilient landscapes [35]. However, re-establishing historical fire patterns in changing landscapes and climates without historical analogues is challenging [36].

Another widely used approach considers species and ecosystem characteristics to define suitable fire patterns [4]. This approach recognizes the context-dependence of burning strategies, varying with ecosystem type, historical fire regime, land management, conservation goals, and societal objectives and constraints [37].

Both of these approaches may involve the application of fire. While fire management spans a continuum of spatio-temporal scales, it is useful to consider two overarching types of planned burning: (i) broadcast burning, applied more uniformly over large areas (from hundreds to thousands of hectares) and emphasizing high coverage of burnt areas across the landscape; and (ii) patch-mosaic burning, conducted more heterogeneously at smaller scales (from less than a hectare to hundreds of hectares) and emphasizing the creation of patchiness and variation in burnt areas. Delineating burning strategies can help predict spatial and temporal fire patterns and their ecological effects. Broadcast burns are resource-intensive and come with risks of fire escaping its intended bounds and purpose. However, these large-scale burns can also be patchy depending on environmental variation (vegetation types, topography or fuel moisture) and ignition patterns, sometimes leaving 30–70% unburnt within a fire perimeter, as seen in examples from Australia [38], South Africa [39] and the western USA [40]. Patch-mosaic burning aligns more closely with traditional fire practices [41]. These two categories blur in real landscapes, where historical, traditional and current fire use often defies easy classification [42].

(b) Biodiversity outcomes of historical and contemporary fire uses

Burning practices create different types of landscape mosaics [43] affecting biodiversity both positively and negatively (see examples in electronic supplementary material, table S1). A common aim is to promote particular vegetation types or species [3] by forming vegetation mosaics encompassing both younger and older successional patches [41]. These practices often involve low- to mixed-severity but frequent fires, set during moderate weather conditions to limit fire spread and intensity [44].

Fire is used to manage open habitats in various regions of the world. In British Columbia, Canada, a common goal is rangeland management, with wildlife benefits often a secondary consideration to production of livestock and pasture [45]. In the USA Great Plains, prescribed burning is used to restore the ecological functionality of grasslands by creating heterogeneous patches through a mix of fire and grazing practices [40]. However, annual burning still occurs over large areas, with some evidence that this results in uniform landscapes and reduced biodiversity [46]. In Southern Africa, a century of fire use in grassy ecosystems shows that modifying fire patterns may help to maintain herbaceous plant diversity and the animals that feed on them [47,48]. In Northern Australia's tropical woodlands, shifting from late dry-season fires, which are often larger and uncontrolled, to early dry-season fires, which as typically smaller and patchier prescribed fires, may protect longer-unburnt habitats and enhance biodiversity by retaining important habitat structures [49,50]. Similarly, in the Brazilian Cerrado, some types of fire help maintain desirable structural and floristic components of savanna landscapes while reducing the occurrence of high-severity dry-season wildfires [51].

Planned fires are also used to manage woodland and forest habitats. In Scandinavian conifer forests, some types of pre-scribed fire promote more 'natural' conditions characterized by heterogeneous stand structures that favour pyrophilous and saproxylic organisms [52]. In many regions there is a need to reconcile wood production with conservation objectives. One way that this is being explored in the south eastern USA is through application of low-intensity fires in combination with silvicultural practices; evidence from stands of *Pinus palustris* indicates that this approach can maintain diverse understories of plants and fauna that benefit from a more open canopy [53]. Fire plays an important role in preserving heathland habitats of high conservation value in the UK and Europe [54]. However, in the UK, for example, planned burning in heathlands dominated by *Calluna vulgaris* primarily aims to maintain vegetation that supports the hunting of *Lagopus lagopus*. This has led to conflict between conservation objectives and the hunting objectives [55], with an ongoing challenge of maintaining fire intervals that promote *Calluna* and graminoids while preserving peat-forming mosses which are more sensitive to fire.

Although fire is also widely used to manage ecosystem services such as food and materials production [56], carbon maintenance [50] and clean water, the use of prescribed fires in many regions is primarily driven by the need to protect people from wildfires [37]. Hence, new frameworks are needed that consider biodiversity alongside other societal values.

(c) Biodiversity-related issues and limitations of fire use

We identify four main challenges in applying fire for biodiversity conservation: (i) setting objectives for biodiversity conservation and planned burning, (ii) the complexity of fire-biodiversity relationships, (iii) uncertainty about past and future fire patterns, and (iv) creating the landscape types that meet biodiversity goals.

Setting ecological burn objectives is challenging due to the need to consider all species involved, not just a select few. A meta-analysis on the effects of prescribed burning effects on biodiversity found difficulties in detecting consistent relationships due to study heterogeneity and insufficient comparability and reporting across studies [57]. Limited information often leads to burning practices that prioritize more easily measured taxa, such as plants [37,58], while animals and their habitats are often neglected [58,59]. Moreover, growing evidence shows that using fire to benefit specific animals, like large savanna herbivores [37], could affect the overall biotic community. Thus, well-defined objectives should consider multiple taxa to achieve desired outcomes [36].

Fire patterns are complex and may threaten biodiversity in fire-adapted systems in a range of ways. In Australia, for example, fires that are too frequent can harm threatened vertebrates that require long-unburnt habitats [60–62]. At the same time, some threatened vertebrates are not getting enough of the 'right' kind of fire [63]. The timing of fires is also important: events outside the peak fire season can reduce flowering and alter seed chemistry of plants stimulated to flower by fire, such as *Doryanthes excelsa* from eastern Australia [64]. Broadcast burning for fuel reduction (approx. 5% of the landscape per year [65]) can lead to excessive juvenile vegetation, decreasing habitat for intermediate and mature seral species [66,67]. Therefore, complexity of fire–biodiversity relationships under a changing climate means that planned burning—whether it be broadcast

burning or patch-mosaic burning—must carefully consider impacts on biodiversity at a range of temporal and spatial scales [68–71].

Another challenge is the uncertainty surrounding fire–biodiversity relationships under novel and emerging conditions. Relying solely on historical fire regimes as reference levels does not guarantee achieving desired ecological outcomes because many ecosystems now harbour new mixes of species, undergo more extreme climates and are subject to different human land-uses [33–35]. Invasion by alien plants, declining habitat quality, or loss of species associated with specific fire regimes are growing risks in many parts of the world [72]. Prescribed burning programmes aimed at achieving large-scale objectives, such as reduction in greenhouse gas emissions, may not necessarily produce local biodiversity co-benefits in fire-adapted ecosystems [50].

Lastly, achieving desirable fire patterns that promote conservation is difficult. While generating diverse or patchy fire patterns provide opportunities to conserve many species, a highly variable burning regime does not necessarily ensure increased biodiversity because desirable fire types, scale of burning and effectiveness are context-specific [73]. Generating adequate fire mosaics, including variation in patch size, connectivity [74] and time-since-fire distribution, is also difficult due to insufficient understanding of the ecological effects of different mosaics, contributing to management strategies that may not be fit for purpose [58]. A better understanding of context-dependence and key mechanisms underpinning fire–biota relationships can improve the creation of desirable forms of fire-driven variation, sometimes called pyrodiversity [75]. However, empirical studies indicate that pyrodiversity–biodiversity relationships are not straightforward [76], varying with climate [77], biota and ecosystems, and influenced by how pyrodiversity is defined and the spatio-temporal scale of analysis [78].

3. Best practices for biodiversity-enhancing fire management

The fast pace of fire-related changes [79], and increasingly novel conditions means that biodiversity management needs to be adaptive. One-size-fits-all approaches that overlook complexity and local context must be revised [80,81]. In the following sections, we describe a framework for enhancing decision-making practices in fire management for biodiversity conservation. We draw from several fields of research, including adaptive management [82,83], structured decision-making [84] and decision science [41]. The proposed framework (figure 1) encompasses the following steps: (i) specifying objectives and indicators for evaluating management alternatives; (ii) developing management alternatives to address the objectives; (iii) analysing potential consequences and considering trade-offs and uncertainties; and (iv) implementing strategies while monitoring their effectiveness, and sharing results to foster collaboration and new knowledge.

(a) Identifying objectives

Effectively protecting biodiversity under novel fire regimes demands knowledge of ecosystems, as well as societal needs. Policymakers and practitioners often use a range of information to formulate fire management objectives, including spatial planning tools, such as habitat and species distribution maps, conservation plans, local knowledge about forest management and historical fire records. Indigenous and traditional fire knowledge, as well as Indigenous-led initiatives, are also increasingly recognized [1]. Engaging a wide range of community stakeholders provides valuable perspectives and helps clarify ambiguities for policymakers on the potential implications of their decisions [84]. It can also help to develop biodiversity objectives that are well-defined and measurable, improve their integration with other fire management objectives such as resource management, protection of lives and properties and supporting cultural practices [85], while also shifting towards allowing natural dynamics for self-organization [86].

Common objectives of ecological fire management include the generation of different types of pyrodiversity and fire mosaics [78], the preservation of threatened species and the restoration of ecosystem processes. More specifically, objectives can range from more precise goals [87], such as increasing food quality for a single species, like the mouflon (*Ovis orientalis*) in France [88], to broader goals, like creating open habitats for multiple species of invertebrates and birds [89,90].

A comprehensive understanding of trade-offs between multiple objectives is needed [41]. Broad scale or patchy burning can achieve multiple goals (figure 2), but landscape fire planning must balance diverse, scale-specific objectives. In the Pyrenees, for example, a strategic fire plan has the objectives of maintaining open rangelands and increasing habitat diversity, and actions to achieve these include smaller planned and larger unplanned fires [108]. After years of land abandonment in the region [109], this kind of prescribed burning is now helping to control shrub encroachment, recover pastoral value and reduce fire hazard. Although habitat improvement for birds of conservation concern is a positive side effect of management in the Pyrenees [56], fires that are too frequent can harm some aspects of biodiversity [110]. This highlights the need for a set of objectives that capture multiple biodiversity and ecosystem values.

Defining management objectives should involve identifying measurable attributes to assess goal fulfilment and adaptation [111,112]. New approaches are helping to do this. For example, fire severity mapping offers insights into vegetation recovery and biological legacies, crucial for understanding fire's effects on biodiversity and ecosystem services. Emerging technologies, such as unoccupied aerial vehicles and acoustic recording devices, are also powerful tools for tracking post-fire changes in animal distributions [113]. Measurement of biodiversity and ecosystem processes enables ongoing monitoring, which facilitates adaptive management by identifying trends and opportunities to improve actions and strategies [114]. It also enhances fire management transparency and credibility by demonstrating the impact of different initiatives, supporting both conservation and socio-economic goals [115].
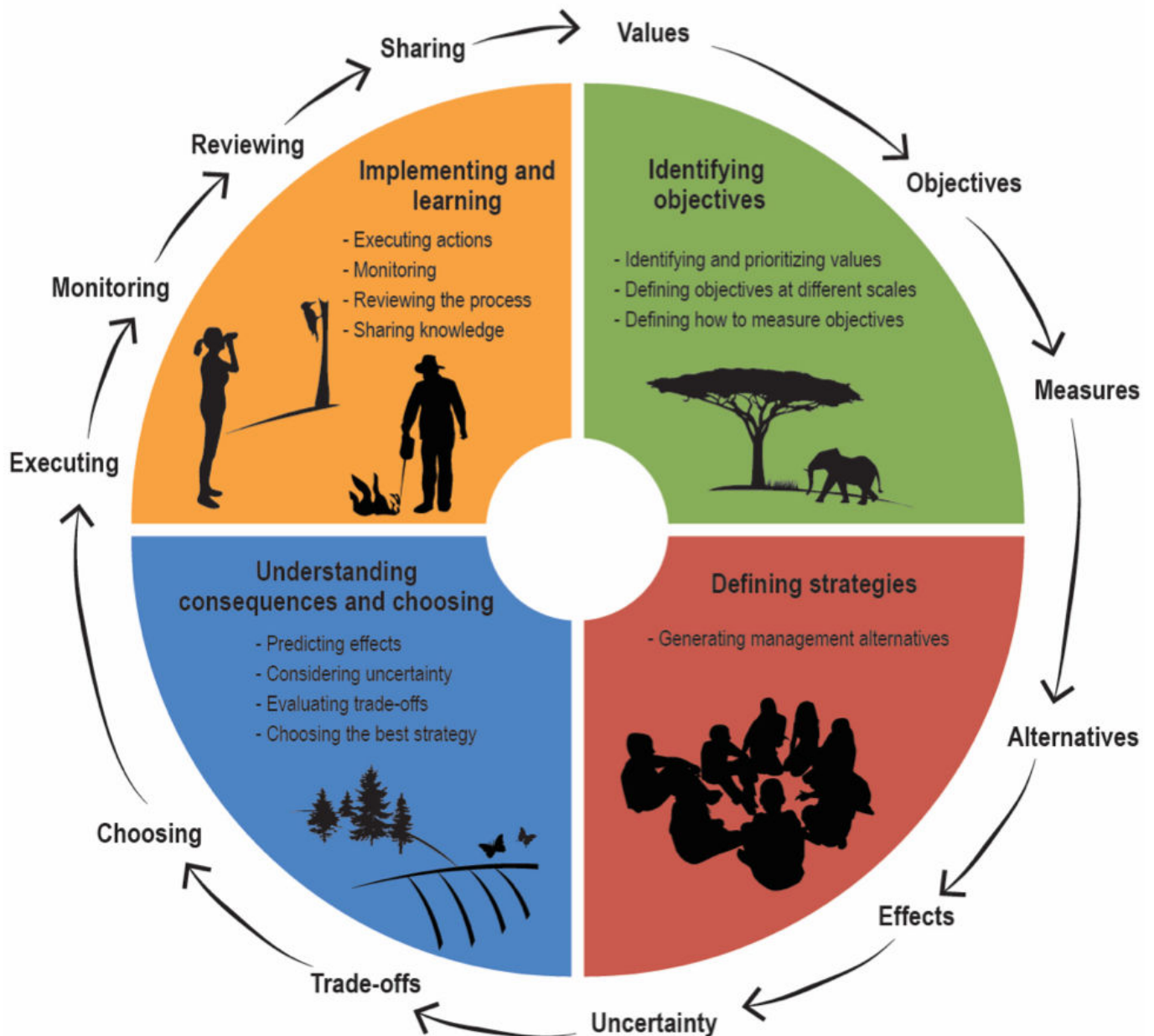
Figure 1. A decision-making framework to implement fire and biodiversity management under novel fire regimes.

(b) Defining strategies

Increased fire activity in many ecosystems experiencing more extreme climatic conditions intensifies the need for strategic intervention to promote biodiversity. A new approach is to use 'adaptation menus'—collections of fire and climate adaptation strategies developed through science–management partnerships [116]. For example, climate change and increased fire activity are transforming montane and subalpine forests into grasslands in the Greater Yellowstone, USA [117,118]. In this region, fire suppression activities are more feasible in subalpine forests but what is on the 'menu' may differ depending on the vegetation type or region. For example, maintaining low fuel loads is essential in drier conifer forests to sustain frequent, low-severity fire regimes, as seen in Wyoming, USA [119].

Incorporating socio-economic aspects into fire management strategies involves acknowledging fire's economic value, including burns for agriculture or traditional land-use [120]. Assessing unintended impacts of fire management on local economies, and ensuring interventions do not inadvertently harm people or assets, is also crucial. By considering socio-economic contexts and involving local stakeholders, fire management can move towards achieving ecological preservation while supporting community well-being and cultural values [121,122].

Globally, there is a wealth of Indigenous, traditional and local expertise in fire management and ecosystem stewardship [1,123,124]. In many fire-prone biomes, such as the Brazilian Cerrado, Indigenous fire practices enhance food availability for small vertebrates and arthropods, increasing species diversity [125]. In Australia, collaboration between Indigenous communities and land management agencies has positively impacted a range of ecological and social values [82,96,126]. Creating learning networks among scientific and non-scientific communities is one of many ways that people can come together to develop strategies that meet the challenges posed by novel fire regimes [127].
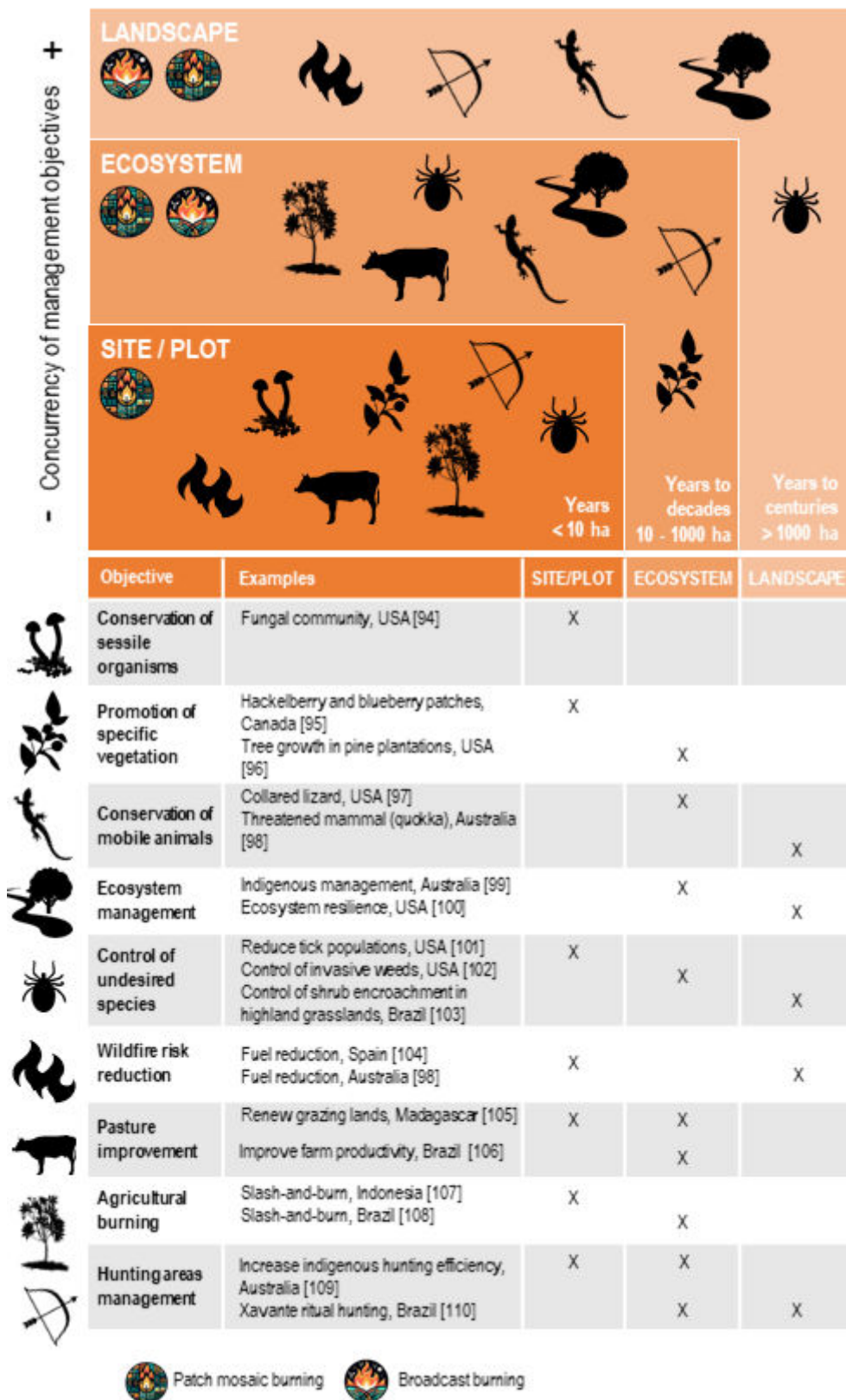
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Objective	Examples	SITE/PLOT	ECOSYSTEM	LANDSCAPE
Conservation of sessile organisms	Fungal community, USA [94]	X		
Promotion of specific vegetation	Hackelberry and blueberry patches, Canada [95]	X		
	Tree growth in pine plantations, USA [96]		X	
Conservation of mobile animals	Collared lizard, USA [97]		X	
	Threatened mammal (quokka), Australia [98]			X
Ecosystem management	Indigenous management, Australia [99]		X	
	Ecosystem resilience, USA [100]			X
Control of undesired species	Reduce tick populations, USA [101]	X		
	Control of invasive weeds, USA [102]		X	
	Control of shrub encroachment in highland grasslands, Brazil [103]			X
Wildfire risk reduction	Fuel reduction, Spain [104]	X		
	Fuel reduction, Australia [98]			X
Pasture improvement	Renew grazing lands, Madagascar [105]	X	X	
	Improve farm productivity, Brazil [106]		X	
Agricultural burning	Slash-and-burn, Indonesia [107]	X		
	Slash-and-burn, Brazil [108]		X	
Hunting areas management	Increase indigenous hunting efficiency, Australia [109]	X	X	
	Xavante ritual hunting, Brazil [110]		X	X

Figure 2. Examples of objectives of fire use aimed at improving biodiversity and other values at different spatial scales. In the attached table under the figure [91–107] provides more information.

(c) Understanding consequences

Analysis of whether or not fire management is working depends on the objectives and the spatial and temporal scales of the strategy (figure 2). This may include considering how a specific fire type affects a species' distribution, population size and movement, or ecosystem structure and function, including metapopulation dynamics [6,127–129]. These effects should be considered at appropriate timescales, with fire initiating ecological changes from hours and days to decades or more [130]. Integrating this knowledge into the complexity and stochasticity of natural systems is necessary. A wide range of approaches

are available to synthesize the consequences of fire management on biodiversity, including those that focus on a range of potential strategies and those more focused on a single best strategy or 'optimal' approach [32,120].

Identifying and quantifying uncertainty relating to the consequences of fire management is important. Under novel fire regimes, in the physiology, phenology, composition and structure of plant communities is likely to be modified [131]. Climate and fire regime changes may render some data and knowledge inadequate for new conditions [132]. To address this, a useful approach for building on 'best practice' knowledge is to integrate species vulnerability assessments, that to date are largely focused on climate change, with new empirical observations [133]. There are exciting opportunities to develop this approach to fire regimes and the resulting shifts in biodiversity [112]. This challenge is particularly evident in ecosystems like South African savannas, where low-intensity grass fires can transition into high-intensity 'firestorms' under specific conditions, exacerbated by climate change [134]. Such extreme conditions alter ecosystem dynamics and complicate the implementation of prescribed burns [135]. Therefore, decision-makers must consider a range of values—including cultural, social and economic factors—when developing strategies to address these uncertainties [136,137].

Once the consequences of management alternatives are estimated, and uncertainties accounted for appropriately, clarifying trade-offs between different strategies is a useful next step [84]. This involves examining how different objectives, and performance measures, are likely to change under alternative fire management strategies [138]. Fire management strategies can harm biodiversity [41]. So, recognizing that there are trade-offs is crucial for informed decisions about biodiversity conservation in the context of dealing with other values such as human health and built assets (Box 1).

Box 1. Examples of trade-offs in fire use and possible ways to navigate them.

Reduce wildfire risks to people versus preserve natural fire regimes: fire suppression to protect human communities can disrupt historical fire regimes [139]. Taking a long-term perspective may help to balance immediate protection of lives and property with ecosystem health.

Improve grazing quality versus protect native biodiversity: fire can benefit grazing animals by removing shrubs but may also spread invasive plants [140]. Understanding this risk, spatial planning may help avoid burning in environments sensitive to plant invasion and grazing animals.

Wildfire prevention versus biodiversity conservation: prescribed burning for hazard reduction can negatively impact biodiversity. A risk-based framework provides opportunities to carefully assess the outcomes of prescribed burning on multiple values while considering critical uncertainties [72].

Fuel management versus tree species composition: prescribed burning in *Pinus yunnanensis* forests benefits some understorey *Quercus* species but harms others taxa [141]. Better knowledge of plant life histories and regeneration capacities aids conservation and fuel management goals.

Carbon storage versus biodiversity conservation: managing forests for *Leuconotopicus borealis* increases biodiversity but decreases carbon storage potential [142]. Trial prescribed burning and thinning to balance species conservation and carbon sequestration objectives.

Traditional pasture management versus natural area protection: restricting fire in a protected area conflicts with traditional grassland renewal by fire, and may negatively affect biodiversity [143]. Understanding the consequences of fire through experimentation with local communities can help to develop conservation objectives, while meeting cultural and economic goals.

(d) Implementing and learning

In the context of novel fire regimes, effective information exchange, planning and preparation are essential for implementing strategies successfully. However, administrative, social, legal, logistic, budgetary and weather-related issues can delay or cancel burning programmes. Flexibility—openness to dialogue, adapting schedules, modifying techniques or redistributing resources—helps overcome these obstacles. For example, with an estimated 17% reduction in the window for prescribed burning in the western United States under 2°C of global warming, adjusting burning periods will be necessary [144], though identifying the optimal window requires tools not always available to managers.

Monitoring should have clear objectives and efficient methods for comparing attributes across fire management scenarios over time. Methods must balance effort (time, personnel, logistics and budget) with quality data. Monitoring, combined with mapping, occurs in phases: before, during and after the burn. Pre-burn data informs the burn plan and establishes a baseline [145]. Data collected during the burn helps refine operations [146]. Immediate fire effects—from structural changes to organism mortality—may not be fully apparent initially [147], but monitoring over weeks to years allows a comprehensive assessment of ecological resilience and management effectiveness. Long-term monitoring is required to assess ecosystem resilience, including vegetation and wildlife recovery [110].

Observational, experimental and modelling approaches are invaluable for predicting outcomes of prescribed fires and integrating this knowledge into planning. BACI (before-after-control-impact) designs are useful for assessing ecological effects of fire [148,149], but they are logistically demanding, expensive and may not yield rapid results [150]. Alternatively, before-after-only or control-impact designs are also useful, provided there is adequate sample size and spatio-temporal replication, and treatment interspersion and synchronic sampling of burnt and unburnt sites [145].

Reviewing progress towards goals is essential. It involves assessing objective achievement, unexpected results, and stakeholder satisfaction, prompting evaluation. Identifying areas for improvement or incorporating new information should lead to adjustments at any stage (figure 1). This process is particularly crucial as the environment rapidly change and uncertainty increases [82].

Effective knowledge dissemination must prioritize stakeholder inclusivity, ensuring their contributions are recognized and they are well-informed about decision implications. This transparency fosters ownership and supports future initiatives [151]. Moreover, transferring new knowledge to policymakers and stakeholders helps reduce ambiguities in future decision-making.

4. Fire management for biodiversity: concluding remarks

Effective fire management under novel fire regimes requires an evidence-based approach. Biodiversity-focused fire management must recognize fire's ecological role, adapting strategies to mimic natural or historical fire regimes while accounting for novel conditions driven by rapid environmental changes. This demands clear objectives, long-term ecological knowledge and robust climate predictions. Evidence-based management, supported by diverse research methods and long-term studies, is essential to assess trade-offs and address knowledge gaps. Monitoring ensures alignment with conservation goals, with feedback mechanisms enabling ongoing adjustments. Effective management integrates fire activities with broader land management and acknowledges fire's ecological and cultural significance.

To ensure fire management for biodiversity conservation can deal with emerging conditions, we suggest the following ways forward: (i) consider both social and ecological factors that influence fire and biodiversity and how they are valued; (ii) establish long-term biodiversity monitoring in fire-affected ecosystems making use of new remote sensing technologies (like UAVs and LiDAR), opportunities for on-ground and real-time observations of ecosystems, to refine strategies; (iii) prioritize cross-disciplinary collaboration that fosters knowledge among and between local communities, fire practitioners, conservation managers and scientists; and (iv) implement iterative testing of fire management actions, supported by adaptive management and experimentation (such as pilot projects), and continual learning. This integrated approach to learning about fire and biodiversity will help to navigate the complexities of novel fire regimes and preserve biodiversity in a fast-changing world.

References

1. Maclean K, Hankins DL, Christianson AC, Oliveras I, Bilbao BA, Costello O, Langer ER, Robinson CJ. 2023 Revitalising indigenous cultural fire practice: benefits and partnerships. *Trends Ecol. Evol.* **38**, 899–902. (doi:10.1016/j.tree.2023.07.001)

2. Pyne SJ. 2021 *The pyrocene: how we created an age of fire, and what happens next*. Berkeley, CA: University of California Press. (doi:10.1525/9780520383593)

3. Weir JR, Scasta JD. 2022 *Global application of prescribed fire*. Clayton, Australia: Csiro Publishing.

4. Kelly LT *et al.* 2020 Fire and biodiversity in the Anthropocene. *Science* **370**, eabb0355. (doi:10.1126/science.abb0355)

5. Rodrigues M, Jiménez-Ruano A, de la Riva J. 2020 Fire regime dynamics in mainland Spain. Part 1: drivers of change. *Sci. Total Environ.* **721**, 135841. (doi:10.1016/j.scitotenv.2019.135841)

6. He T, Lamont BB, Pausas JG. 2019 Fire as a key driver of Earth's biodiversity. *Biol. Rev.* **94**, 1983–2010. (doi:10.1111/brv.12544)

7. Wu C, Sitch S, Huntingford C, Mercado LM, Venevsky S, Lasslop G, Archibald S, Staver AC. 2022 Reduced global fire activity due to human demography slows global warming by enhanced land carbon uptake. *Proc. Natl Acad. Sci. USA* **119**, e2101186119. (doi:10.1073/pnas.2101186119)

8. Ellis EC. 2021 Land use and ecological change: a 12,000-year history. *Annu. Rev. Environ. Resour*. **46**, 1–33. (doi:10.1146/annurev-environ-012220-010822)

9. Hughes TP, Carpenter S, Rockström J, Scheffer M, Walker B. 2013 Multiscale regime shifts and planetary boundaries. *Trends Ecol. Evol*. **28**, 389–395. (doi:10.1016/j.tree.2013.05.019)

10. Andela N *et al*. 2017 A human-driven decline in global burned area. *Science* **356**, 1356–1362. (doi:10.1126/science.aal4108)

11. Mouillot F, Field CB. 2005 Fire history and the global carbon budget: a 1°× 1° fire history reconstruction for the 20th century. *Glob. Chang. Biol*. **11**, 398–420. (doi:10.1111/j.1365-2486.2005.00920.x)

12. Kelly LT, Fletcher MS, Oliveras Menor I, Pellegrini AFA, Plumanns-Pouton ES, Pons P, Williamson GJ, Bowman DMJS. 2023 Understanding fire regimes for a better Anthropocene. *Annu. Rev. Environ. Resour*. **48**, 207–235. (doi:10.1146/annurev-environ-120220-055357)

13. Archibald S, Hempson GP, Lehmann C. 2019 A unified framework for plant life-history strategies shaped by fire and herbivory. *New Phytol*. **224**, 1490–1503. (doi:10.1111/nph.15986)

14. Harrison SP *et al*. 2021 Understanding and modelling wildfire regimes: an ecological perspective. *Environ. Res. Lett*. **16**, 125008. (doi:10.1088/1748-9326/ac39be)

15. Lamont BB, Pausas JG, He T, Witkowski ETF, Hanley ME. 2020 Fire as a selective agent for both serotiny and nonserotiny over space and time. *Crit. Rev. Plant Sci*. **39**, 140–172. (doi:10.1080/07352689.2020.1768465)

16. Moreira B, Tormo J, Estrelles E, Pausas JG. 2010 Disentangling the role of heat and smoke as germination cues in Mediterranean Basin flora. *Ann. Bot*. **105**, 627–635. (doi:10.1093/aob/mcq017)

17. Shen Y, Cai W, Prentice IC, Harrison SP. 2023 Community abundance of resprouting in woody plants reflects fire return time, intensity, and type. *Forests* **14**, 878. (doi:10.3390/f14050878)

18. Charles-Dominique T, Midgley GF, Tomlinson KW, Bond WJ. 2018 Steal the light: shade vs fire adapted vegetation in forest–savanna mosaics. *New Phytol*. **218**, 1419–1429. (doi:10.1111/nph.15117)

19. Viljur ML *et al*. 2022 The effect of natural disturbances on forest biodiversity: an ecological synthesis. *Biol. Rev*. **97**, 1930–1947. (doi:10.1111/brv.12876)

20. Puig-Gironès R, Santos X, Bros V. 2023 Temporal differences in snail diversity responses to wildfires and salvage logging. *Environ. Conserv*. **50**, 40–49. (doi:10.1017/s0376892922000443)

21. Regos A, D'Amen M, Titeux N, Herrando S, Guisan A, Brotons L. 2016 Predicting the future effectiveness of protected areas for bird conservation in Mediterranean ecosystems under climate change and novel fire regime scenarios. *Divers. Distrib*. **22**, 83–96. (doi:10.1111/ddi.12375)

22. van Nes EH, Staal A, Hantson S, Holmgren M, Pueyo S, Bernardi RE, Flores BM, Xu C, Scheffer M. 2018 Fire forbids fifty-fifty forest. *PLoS One* **13**, e0191027. (doi:10.1371/journal.pone.0191027)

23. Lasslop G *et al*. 2020 Global ecosystems and fire: multi-model assessment of fire-induced tree-cover and carbon storage reduction. *Glob. Chang. Biol*. **26**, 5027–5041. (doi:10.1111/gcb.15160)

24. Pellegrini AFA *et al*. 2021 Decadal changes in fire frequencies shift tree communities and functional traits. *Nat. Ecol. Evol*. **5**, 504–512. (doi:10.1038/s41559-021-01401-7)

25. Moreira F *et al*. 2020 Wildfire management in Mediterranean-type regions: paradigm change needed. *Environ. Res. Lett*. **15**, 011001. (doi:10.1088/1748-9326/ab541e)

26. Dupuy JL, Fargeon H, Martin-StPaul N, Pimont F, Ruffault J, Guijarro M, Hernando C, Madrigal J, Fernandes P. 2020 Climate change impact on future wildfire danger and activity in southern Europe: a review. *Ann. For. Sci*. **77**, 1–24. (doi:10.1007/s13595-020-00933-5)

27. Sharma S *et al*. 2022 North American tree migration paced by climate in the West, lagging in the East. *Proc. Natl Acad. Sci. USA* **119**, e2116691118. (doi:10.1073/pnas.2116691118)

28. Mason DS, Lashley MA. 2021 Spatial scale in prescribed fire regimes: an understudied aspect in conservation with examples from the southeastern United States. *Fire Ecol*. **17**, 3. (doi:10.1186/s42408-020-00087-9)

29. Bradshaw SD, Dixon KW, Lambers H, Cross AT, Bailey J, Hopper SD. 2018 Understanding the long-term impact of prescribed burning in mediterranean-climate biodiversity hotspots, with a focus on south-western Australia. *Int. J. Wildland Fire* **27**, 643–657. (doi:10.1071/WF18067)

30. Brotons L, Duane A. 2019 Correspondence: uncertainty in climate-vegetation feedbacks on fire regimes challenges reliable long-term projections of burnt area from correlative models. *Fire* **2**, 8. (doi:10.3390/fire2010008)

31. Costafreda-Aumedes S, Comas C, Vega-Garcia C. 2017 Human-caused fire occurrence modelling in perspective: a review. *Int. J. Wildland Fire* **26**, 983–998. (doi:10.1071/wf17026)

32. Pais S, Aquilué N, Brotons L, Honrado JP, Fernandes PM, Regos A. 2023 The REMAINS R-package: paving the way for fire-landscape modeling and management. *Environ. Model. Softw*. **168**, 105801. (doi:10.1016/j.envsoft.2023.105801)

33. Lashley MA, Chitwood MC, Prince A, Elfelt MB, Kilburg EL, DePerno CS, Moorman CE. 2014 Subtle effects of a managed fire regime: a case study in the longleaf pine ecosystem. *Ecol. Indic*. **38**, 212–217. (doi:10.1016/j.ecolind.2013.11.006)

34. Freeman J, Kobziar L, Rose EW, Cropper W. 2017 A critique of the historical-fire-regime concept in conservation. *Conserv. Biol*. **31**, 976–985. (doi:10.1111/cobi.12942)

35. Moritz MA, Hurteau MD, Suding KN, D'Antonio CM. 2013 Bounded ranges of variation as a framework for future conservation and fire management. *Ann. NY Acad. Sci*. **1286**, 92–107. (doi:10.1111/nyas.12104)

36. Penman TD *et al*. 2011 Prescribed burning: how can it work to conserve the things we value? *Int. J. Wildland Fire* **20**, 721–733. (doi:10.1071/wf09131)

37. van Wilgen BW. 2013 Fire management in species-rich Cape fynbos shrublands. *Front. Ecol. Environ*. **11**, e35–e44. (doi:10.1890/120137)

38. Sitters H, Di Stefano J, Christie FJ, Sunnucks P, York A. 2015 Bird diversity increases after patchy prescribed fire: implications from a before–after control–impact study. *Int. J. Wildland Fire* **24**, 690–701. (doi:10.1071/wf14123)

39. van Wilgen BW, Govender N, Smit IPJ, MacFadyen S. 2014 The ongoing development of a pragmatic and adaptive fire management policy in a large African savanna protected area. *J. Environ. Manag*. **132**, 358–368. (doi:10.1016/j.jenvman.2013.11.003)

40. Castro Rego F, Morgan P, Fernandes P, Hoffman C. 2021 Integrated fire management. In *Fire science: from chemistry to landscape management* (eds FC Rego, P Morgan, P Fernandes, C Hoffman), pp. 509–597. Cham, Switzerland: Springer International Publishing. (doi:10.1007/978-3-030-69815-7_13)

41. Driscoll DA *et al*. 2010 Resolving conflicts in fire management using decision theory: asset-protection versus biodiversity conservation. *Conserv. Lett*. **3**, 215–223. (doi:10.1111/j.1755-263x.2010.00115.x)

42. Pivello VR *et al*. 2021 Understanding Brazil's catastrophic fires: causes, consequences and policy needed to prevent future tragedies. *Perspect. Ecol. Conserv*. **19**, 233–255. (doi:10.1016/j.pecon.2021.06.005)

43. Trauernicht C, Brook BW, Murphy BP, Williamson GJ, Bowman DMJS. 2015 Local and global pyrogeographic evidence that indigenous fire management creates pyrodiversity. *Ecol. Evol*. **5**, 1908–1918. (doi:10.1002/ece3.1494)

44. Hoffman KM *et al*. 2021 Conservation of Earth's biodiversity is embedded in Indigenous fire stewardship. *Proc. Natl Acad. Sci. USA* **118**, e2105073118. (doi:10.1073/pnas.2105073118)

45. Leverkus SER, Fuhlendorf SD, Geertsema M, Engle DM. 2022 Fire in the boreal forest, north-east British Columbia, Canada: putting fire out on the land. In *Global application of prescribed fire* (eds JR Weir, JD Scasta), pp. 14–37. Clayton, Australia: CSIRO Publishing.

46. Hamilton RG. 2022 An NGO and prescribed fire: the nature conservancy's Joseph H. Williams tallgrass prairie preserve (USA). In *Global application of prescribed fire* (eds JR Weir, JD Scasta), pp. 162–175. Clayton, Australia: CSIRO Publishing.

47. Govender N, Staver C, Archibald S, Wigley-Coetsee C, Strydom T, Humphrey G, Kimuyu D. 2022 Lessons from a century of evidence-based fire management in grassy ecosystems. *Afr. J. Range Forage Sci.* **39**, v–vii. (doi:10.2989/10220119.2022.2035489)

48. Smit IPJ, Asner GP, Govender N, Kennedy-Bowdoin T, Knapp DE, Jacobson J. 2010 Effects of fire on woody vegetation structure in African savanna. *Ecol. Appl*. **20**, 1865–1875. (doi:10.1890/09-0929.1)

49. Evans J, Russell-Smith J. 2019 Delivering effective savanna fire management for defined biodiversity conservation outcomes: an Arnhem Land case study. *Int. J. Wildland Fire* **29**, 386–400. (doi:10.1071/WF18126)

50. Perry JJ, Vanderduys EP, Kutt AS. 2016 Shifting fire regimes from late to early dry-season fires to abate greenhouse emissions does not completely equate with terrestrial vertebrate biodiversity co-benefits on Cape York Peninsula, Australia. *Int. J. Wildland Fire* **25**, 742–752. (doi:10.1071/wf15133)

51. Berlinck CN, Batista EKL. 2020 Good fire, bad fire: it depends on who burns. *Flora* **268**, 151610. (doi:10.1016/j.flora.2020.151610)

52. Halme P *et al*. 2013 Challenges of ecological restoration: lessons from forests in northern Europe. *Biol. Conserv*. **167**, 248–256. (doi:10.1016/j.biocon.2013.08.029)

53. Mitchell RJ, Hiers JK, O'Brien JJ, Jack SB, Engstrom RT. 2006 Silviculture that sustains: the nexus between silviculture, frequent prescribed fire, and conservation of biodiversity in longleaf pine forests of the southeastern United States. *Can. J. For. Res*. **36**, 2724–2736. (doi:10.1139/x06-100)

54. Davies GM, Vandvik V, Marrs R, Velle LG. 2022 Fire management in heather-dominated heaths and moorlands of North-West Europe. In *In global application of prescribed fire* (eds JR Weir, JD Scasta), pp. 194–229. Clayton, Australia: CSIRO Publishing.

55. Davies GM *et al*. 2016 The role of fire in UK peatland and moorland management: the need for informed, unbiased debate. *Phil. Trans. R. Soc. B* **371**, 20150342. (doi:10.1098/rstb.2015.0342)

56. Pons P, Lambert B, Rigolot E, Prodon R. 2003 The effects of grassland management using fire on habitat occupancy and conservation of birds in a mosaic landscape. *Biodivers. Conserv*. **12**, 1843–1860. (doi:10.1023/A:1024191814560)

57. Eales J, Haddaway NR, Bernes C, Cooke SJ, Jonsson BG, Kouki J, Petrokofsky G, Taylor JJ. 2018 What is the effect of prescribed burning in temperate and boreal forest on biodiversity, beyond pyrophilous and saproxylic species? A systematic review. *Environ. Evid*. **7**, 19. (doi:10.1186/s13750-018-0131-5)

58. Clarke MF. 2008 Catering for the needs of fauna in fire management: science or just wishful thinking? *Wildl. Res*. **35**, 385–394. (doi:10.1071/WR07137)

59. Croft P, Hunter JT, Reid N. 2016 Forgotten fauna: habitat attributes of long-unburnt open forests and woodlands dictate a rethink of fire management theory and practice. *For. Ecol. Manag*. **366**, 166–174. (doi:10.1016/j.foreco.2016.02.015)

60. von Takach B, Jolly CJ, Dixon KM, Penton CE, Doherty TS, Banks SC. 2022 Long-unburnt habitat is critical for the conservation of threatened vertebrates across Australia. *Landsc. Ecol*. **37**, 1469–1482. (doi:10.1007/s10980-022-01427-7)

61. Woinarski JCZ, Burbidge AA, Harrison PL. 2015 Ongoing unraveling of a continental fauna: decline and extinction of Australian mammals since European settlement. *Proc. Natl Acad. Sci. USA* **112**, 4531–4540. (doi:10.1073/pnas.1417301112)

62. Driscoll DA *et al*. 2024 Biodiversity impacts of the 2019-2020 Australian megafires. *Nature New Biol*. **635**, 898–905. (doi:10.1038/s41586-024-08174-6)

63. Santos JL, Hradsky BA, Keith DA, Rowe KC, Senior KL, Sitters H, Kelly LT. 2022 Beyond inappropriate fire regimes: a synthesis of fire-driven declines of threatened mammals in Australia. *Conserv. Lett*. **15**, e12905. (doi:10.1111/conl.12905)

64. Paroissien R, Ooi MKJ. 2021 Effects of fire season on the reproductive success of the post-fire flowerer *Doryanthes excelsa*. *Environ. Exp. Bot*. **192**, 104634. (doi:10.1016/j.envexpbot.2021.104634)

65. Fernandes PM. 2015 Empirical support for the use of prescribed burning as a fuel treatment. *Curr For. Rep* **1**, 118–127. (doi:10.1007/s40725-015-0010-z)

66. Puig-Gironès R, Brotons L, Pons P, Franch M. 2023 Examining the temporal effects of wildfires on forest birds: should I stay or should I go? *For. Ecol. Manag*. **549**, 121439. (doi:10.1016/j.foreco.2023.121439)

67. Connell J, Watson SJ, Taylor RS, Avitabile SC, Schedvin N, Schneider K, Clarke MF. 2019 Future fire scenarios: predicting the effect of fire management strategies on the trajectory of high-quality habitat for threatened species. *Biol. Conserv*. **232**, 131–141. (doi:10.1016/j.biocon.2019.02.004)

68. Duane A, Aquilué N, Canelles Q, Morán-Ordoñez A, De Cáceres M, Brotons L. 2019 Adapting prescribed burns to future climate change in Mediterranean landscapes. *Sci. Total Environ*. **677**, 68–83. (doi:10.1016/j.scitotenv.2019.04.348)

69. Holland GJ, Clarke MF, Bennett AF. 2017 Prescribed burning consumes key forest structural components: implications for landscape heterogeneity. *Ecol. Appl*. **27**, 845–858. (doi:10.1002/eap.1488)

70. Bowman DMJS, Borchers-Arriagada N, Macintosh A, Butler DW, Williamson GJ, Johnston FH. 2024 Climate change must be factored into savanna carbon-management projects to avoid maladaptation: the case of worsening air pollution in western Top End of the Northern Territory, Australia. *Rangel. J*. **46**, RJ23049. (doi:10.1071/rj23049)

71. Furlaud JM, Williamson GJ, Bowman DMJS. 2023 Mechanical treatments and prescribed burning can reintroduce low-severity fire in southern Australian temperate sclerophyll forests. *J. Environ. Manag*. **344**, 118301. (doi:10.1016/j.jenvman.2023.118301)

72. Howard T, Burrows N, Smith T, Daniel G, McCaw L. 2020 A framework for prioritising prescribed burning on public land in Western Australia. *Int. J. Wildland Fire* **29**, 314–325. (doi:10.1071/WF19029)

73. Kelly LT, Brotons L. 2017 Using fire to promote biodiversity. *Science* **355**, 1264–1265. (doi:10.1126/science.aam7672)

74. Brotons L, Pons P, Herrando S. 2005 Colonization of dynamic Mediterranean landscapes: where do birds come from after fire? *J. Biogeogr*. **32**, 789–798. (doi:10.1111/j.1365-2699.2004.01195.x)

75. Kelly LT, Brotons L, McCarthy MA. 2017 Putting pyrodiversity to work for animal conservation. *Conserv. Biol*. **31**, 952–955. (doi:10.1111/cobi.12861)

76. Pastro LA, Dickman CR, Letnic M. 2011 Burning for biodiversity or burning biodiversity? Prescribed burn vs. wildfire impacts on plants, lizards, and mammals. *Ecol. Appl*. **21**, 3238–3253. (doi:10.1890/10-2351.1)

77. Beale CM *et al*. 2018 Pyrodiversity interacts with rainfall to increase bird and mammal richness in African savannas. *Ecol. Lett*. **21**, 557–567. (doi:10.1111/ele.12921)

78. Jones GM, Tingley MW. 2022 Pyrodiversity and biodiversity: a history, synthesis, and outlook. *Divers. Distrib*. **28**, 386–403. (doi:10.1111/ddi.13280)

79. Harfoot MBJ *et al*. 2021 Using the IUCN Red List to map threats to terrestrial vertebrates at global scale. *Nat. Ecol. Evol*. **5**, 1510–1519. (doi:10.1038/s41559-021-01542-9)

80. Barlow J, Berenguer E, Carmenta R, França F. 2020 Clarifying Amazonia's burning crisis. *Glob. Chang. Biol*. **26**, 319–321. (doi:10.1111/gcb.14872)

81.   Uyttewaal K, Prat-Guitart N, Ludwig F, Kroeze C, Langer ER. 2023 Territories in transition: how social contexts influence wildland fire adaptive capacity in rural Northwestern European Mediterranean areas. *Fire Ecol.* **19**, 13. (doi:10.1186/s42408-023-00168-5)

82.   Gillson L, Biggs H, Smit IPJ, Virah-Sawmy M, Rogers K. 2019 Finding common ground between adaptive management and evidence-based approaches to biodiversity conservation. *Trends Ecol. Evol.* **34**, 31–44. (doi:10.1016/j.tree.2018.10.003)

83.   Ascoli D, Beghin R, Ceccato R, Gorlier A, Lombardi G, Lonati M, Marzano R, Bovio G, Cavallero A. 2008 Developing an adaptive management approach to prescribed burning: a long-term heathland conservation experiment in north-west Italy. *Int. J. Wildland Fire* **18**, 727–735. (doi:10.1071/WF07114)

84.   Gregory R, Arvai J, Gerber LR. 2013 Structuring decisions for managing threatened and endangered species in a changing climate. *Conserv. Biol.* **27**, 1212–1221. (doi:10.1111/cobi.12165)

85.   Millington JDA, Perkins O, Smith C. 2022 Human fire use and management: a global database of anthropogenic fire impacts for modelling. *Fire* **5**, 87. (doi:10.3390/fire5040087)

86.   Fath BD, Müller F. 2010 Long-term ecosystem dynamics: theoretical concepts of environmental change. In *Long-term ecological research: between theory and application* (eds F Müller, C Baessler, H Schubert, S Klotz), pp. 27–38. Dordrecht, Netherlands: Springer Netherlands. (doi:10.1007/978-90-481-8782-9_3)

87.   Giljohann KM, McCarthy MA, Kelly LT, Regan TJ. 2015 Choice of biodiversity index drives optimal fire management decisions. *Ecol. Appl.* **25**, 264–277. (doi:10.1890/14-0257.1)

88.   Cazau M, Garel M, Maillard D. 2011 Responses of heather moorland and Mediterranean mouflon foraging to prescribed-burning and cutting. *J. Wildl. Manag.* **75**, 967–972. (doi:10.1002/jwmg.117)

89.   Nakas G, Kantsa A, Vujić A, Mescher MC, De Moraes CM, Petanidou T. 2023 Recent fire in a Mediterranean ecosystem strengthens hoverfly populations and their interaction networks with plants. *Ecol. Evol.* **13**, e9803. (doi:10.1002/ece3.9803)

90.   Puig-Gironès R, Brotons L, Pons P. 2022 Aridity, fire severity and proximity of populations affect the temporal responses of open-habitat birds to wildfires. *Biol. Conserv.* **272**, 109661. (doi:10.1016/j.biocon.2022.109661)

91.   Oliver AK, Callaham MA, Jumpponen A. 2015 Soil fungal communities respond compositionally to recurring frequent prescribed burning in a managed southeastern US forest ecosystem. *For. Ecol. Manage.* **345**, 1–9. (doi:10.1016/j.foreco.2015.02.020)

92.   Gottesfeld LMJ. 1994 Aboriginal burning for vegetation management in northwest British Columbia. *Hum. Ecol.* **22**, 171–188. (doi:10.1007/bf02169038)

93.   Iglay RB, Leopold BD, Miller DA, Wes Burger L. 2010 Effect of plant community composition on plant response to fire and herbicide treatments. *For. Ecol. Manage.* **260**, 543–548. (doi:10.1016/j.foreco.2010.05.010)

94.   Neuwald JL, Templeton AR. 2013 Genetic restoration in the eastern collared lizard under prescribed woodland burning. *Mol. Ecol.* **22**, 3666–3679. (doi:10.1111/mec.12306)

95.   Bain K, Wayne A, Bencini R. 2016 Prescribed burning as a conservation tool for management of habitat for threatened species: the quokka, *Setonix brachyurus*, in the southern forests of Western Australia. *Int. J. Wildland Fire* **25**, 608–617. (doi:10.1071/WF15138)

96.   Mariani M *et al*. 2022 Disruption of cultural burning promotes shrub encroachment and unprecedented wildfires. *Front. Ecol. Environ.* **20**, 292–300. (doi:10.1002/fee.2395)

97.   Wollstein K, Creutzburg MK, Dunn C, Johnson DD, O'Connor C, Boyd CS. 2022 Toward integrated fire management to promote ecosystem resilience. *Rangelands* **44**, 227–234. (doi:10.1016/j.rala.2022.01.001)

98.   Gleim ER, Conner LM, Berghaus RD, Levin ML, Zemtsova GE, Yabsley MJ. 2014 The phenology of ticks and the effects of long-term prescribed burning on tick population dynamics in southwestern Georgia and northwestern Florida. *PLoS One* **9**, e112174. (doi:10.1371/journal.pone.0112174)

99.   Ditomaso JM, Brooks ML, Allen EB, Minnich R, Rice PM, Kyser GB. 2006 Control of invasive weeds with prescribed burning. *Weed Technol.* **20**, 535–548. (doi:10.1614/wt-05-086r1.1)

100.   Sühs RB, Giehl ELH, Peroni N. 2020 Preventing traditional management can cause grassland loss within 30 years in southern Brazil. *Sci. Rep.* **10**, 783. (doi:10.1038/s41598-020-57564-z)

101.   Valor T, González-Olabarria JR, Piqué M. 2015 Assessing the impact of prescribed burning on the growth of European pines. *For. Ecol. Manag.* **343**, 101–109. (doi:10.1016/j.foreco.2015.02.002)

102.   Bloesch U. 1999 Fire as a tool in the management of a savanna/dry forest reserve in Madagascar. *Appl. Veg. Sci.* **2**, 117–124. (doi:10.2307/1478888)

103.   Brunel M, Rammig A, Furquim F, Overbeck G, Barbosa HMJ, Thonicke K, Rolinski S. 2021 When do farmers burn pasture in Brazil: a model-based approach to determine burning date. *Rangel. Ecol. Manag.* **79**, 110–125. (doi:10.1016/j.rama.2021.08.003)

104.   Ketterings QM, Tri Wibowo T, van Noordwijk M, Penot E. 1999 Farmers' perspectives on slash-and-burn as a land clearing method for small-scale rubber producers in Sepunggur, Jambi Province, Sumatra, Indonesia. *For. Ecol. Manag.* **120**, 157–169. (doi:10.1016/s0378-1127(98)00532-5)

105.   Fujisaka S *et al*. 1996 Slash-and-burn agriculture, conversion to pasture, and deforestation in two Brazilian Amazon colonies. *Agric. Ecosyst. Environ.* **59**, 115–130. (doi:10.1016/0167-8809(96)01015-8)

106.   Bird DW, Bird RB, Parker CH. 2005 Aboriginal burning regimes and hunting strategies in Australia's Western Desert. *Hum. Ecol.* **33**, 443–464. (doi:10.1007/s10745-005-5155-0)

107.   Welch JR. 2014 Xavante ritual hunting: anthropogenic fire, reciprocity, and collective landscape management in the Brazilian Cerrado. *Hum. Ecol.* **42**, 47–59. (doi:10.1007/s10745-013-9637-1)

108.   Duane A, Oliveres J, Castellarnau X, Pagés J, Bosch M, Arjó G, Rodriguez L, Castellnou M, Brotons L. 2022 Un nuevo plan de manejo del fuego en la val d'Aran (pirineos) basado en la coexistencia con el fuego: detalles de la modelización científica para la toma de decisiones. In *8° Congr. Forestal Español, Lleida, Spain*. Madrid, Spain: Spanish Society of Forestry Sciences.

109.   Roura-Pascual N, Pons P, Etienne M, Lambert B. 2005 Transformation of a rural landscape in the Eastern Pyrenees between 1953 and 2000. *Mt. Res. Dev.* **25**, 252–261. (doi:10.1659/0276-4741(2005)025[0252:TOARLI]2.0.CO;2)

110.   Pons P, Clavero M. 2010 Bird responses to fire severity and time since fire in managed mountain rangelands. *Anim. Conserv.* **13**, 294–305. (doi:10.1111/j.1469-1795.2009.00337.x)

111.   McLauchlan KK *et al*. 2020 Fire as a fundamental ecological process: research advances and frontiers. *J. Ecol.* **108**, 2047–2069. (doi:10.1111/1365-2745.13403)

112.   Giorgis MA *et al*. 2021 A review of fire effects across South American ecosystems: the role of climate and time since fire. *Fire Ecol.* **17**, 1–20. (doi:10.1186/s42408-021-00100-9)

113.   Boone WW, Bankovich BA, Reichert BE, Watson MB, McCleery RA. 2024 Frequent prescribed burns reduce mammalian species richness and occurrence in longleaf pine sandhills. *For. Ecol. Manag.* **553**, 121596. (doi:10.1016/j.foreco.2023.121596)

114.   Puig-Gironès R, Real J. 2022 A comprehensive but practical methodology for selecting biological indicators for long-term monitoring. *PLoS One* **17**, e0265246. (doi:10.1371/journal.pone.0265246)

115.   Trane M, Marelli L, Siragusa A, Pollo R, Lombardi P. 2023 Progress by research to achieve the sustainable development goals in the EU: a systematic literature review. *Sustainability* **15**, 7055. (doi:10.3390/su15097055)

116.   Sample M *et al*. 2022 Adaptation strategies and approaches for managing fire in a changing climate. *Climate* **10**, 58. (doi:10.3390/cli10040058)

117. Abatzoglou JT, Williams AP. 2016 Impact of anthropogenic climate change on wildfire across western US forests. *Proc. Natl Acad. Sci. USA* **113**, 11770–11775. (doi:10.1073/pnas.1607171113)

118. Rammer W, Braziunas KH, Hansen WD, Ratajczak Z, Westerling AL, Turner MG, Seidl R. 2021 Widespread regeneration failure in forests of Greater Yellowstone under scenarios of future climate and fire. *Glob. Chang. Biol.* **27**, 4339–4351. (doi:10.1111/gcb.15726)

119. Hansen WD, Abendroth D, Rammer W, Seidl R, Turner MG. 2020 Can wildland fire management alter 21st-century subalpine fire and forests in Grand Teton National Park, Wyoming, USA? *Ecol. Appl.* **30**, e02030. (doi:10.1002/eap.2030)

120. Lecina-Diaz J, Chas-Amil ML, Aquilué N, Sil Â, Brotons L, Regos A, Touza J. 2023 Incorporating fire-smartness into agricultural policies reduces suppression costs and ecosystem services damages from wildfires. *J. Environ. Manag.* **337**, 117707. (doi:10.1016/j.jenvman.2023.117707)

121. Ponce-Calderón LP, Limón-Aguirre F, Rodríguez I, Rodríguez-Trejo DA, Bilbao BA, Álvarez-Gordillo G del C, Villanueva-Díaz J. 2022 Fire management in pyrobiocultural landscapes, Chiapas, Mexico. *Trop. For. Issues* **61**, 53–59. (doi:10.55515/ABWJ7126)

122. Otero I, Castellnou M, González I, Arilla E, Castell L, Castellví J, Sánchez F, Nielsen JØ. 2018 Democratizing wildfire strategies. Do you realize what it means? Insights from a participatory process in the Montseny region (Catalonia, Spain). *PLoS One* **13**, e0204806. (doi:10.1371/journal.pone.0204806)

123. Mistry J, Bilbao BA, Berardi A. 2016 Community owned solutions for fire management in tropical ecosystems: case studies from Indigenous communities of South America. *Phil. Trans. R. Soc. B* **371**, 20150174. (doi:10.1098/rstb.2015.0174)

124. Mistry J, Schmidt IB, Eloy L, Bilbao B. 2019 New perspectives in fire management in South American savannas: the importance of intercultural governance. *Ambio* **48**, 172–179. (doi:10.1007/s13280-018-1054-7)

125. Welch JR, Brondízio ES, Hetrick SS, Coimbra CEA. 2013 Indigenous burning as conservation practice: Neotropical savanna recovery amid agribusiness deforestation in Central Brazil. *PLoS One* **8**, e81226. (doi:10.1371/journal.pone.0081226)

126. Bliege Bird R, Codding BF, Kauhanen PG, Bird DW. 2012 Aboriginal hunting buffers climate-driven fire-size variability in Australia's spinifex grasslands. *Proc. Natl Acad. Sci. USA* **109**, 10287–10292. (doi:10.1073/pnas.1204585109)

127. Campbell CB, Blair S, Wilson A. 2010 *Adaptive management of fire: the role of a learning network*. Melbourne, Australia: Victorian Government Department of Sustainability and Environment.

128. Zhang H, Bearup D, Nijs I, Wang S, Barabás G, Tao Y, Liao J. 2021 Dispersal network heterogeneity promotes species coexistence in hierarchical competitive communities. *Ecol. Lett.* **24**, 50–59. (doi:10.1111/ele.13619)

129. Cianciaruso MV, Silva IA, Batalha MA, Gaston KJ, Petchey OL. 2012 The influence of fire on phylogenetic and functional structure of woody savannas: moving from species to individuals. *Perspect. Plant Ecol. Evol. Syst.* **14**, 205–216. (doi:10.1016/j.ppees.2011.11.004)

130. Connell JH, Slatyer RO. 1977 Mechanisms of succession in natural communities and their role in community stability and organization. *Am. Nat.* **111**, 1119–1144. (doi:10.1086/283241)

131. Barbosa JPRAD, Rambal S, Soares AM, Mouillot F, Nogueira JMP, Martins GA. 2012 Plant physiological ecology and the global changes. *Ciênc. Agrotecnol* **36**, 253–269. (doi:10.1590/s1413-70542012000300001)

132. Yousefpour R, Hanewinkel M. 2016 Climate change and decision-making under uncertainty. *Curr. For. Rep.* **2**, 143–149. (doi:10.1007/s40725-016-0035-y)

133. Watson JEM, Rao M, Ai-Li K, Yan X. 2012 Climate change adaptation planning for biodiversity conservation: a review. *Adv. Clim. Chang. Res* **3**, 1–11. (doi:10.3724/sp.j.1248.2012.00001)

134. Browne C, Bond W. 2011 Firestorms in savanna and forest ecosystems: curse or cure? *Veld Flora* **97**, 62–63.

135. Beckett H, Staver AC, Charles-Dominique T, Bond WJ. 2022 Pathways of savannization in a mesic African savanna–forest mosaic following an extreme fire. *J. Ecol.* **110**, 902–915. (doi:10.1111/1365-2745.13851)

136. Wardekker JA. 2011 Climate change impact assessment and adaptation under uncertainty. PhD thesis, Utrecht University, Utrecht, The Netherlands.

137. Taylor B, Chapron G, Kopnina H, Orlikowska E, Gray J, Piccolo JJ. 2020 The need for ecocentrism in biodiversity conservation. *Conserv. Biol.* **34**, 1089–1096. (doi:10.1111/cobi.13541)

138. Driscoll DA, Bode M, Bradstock RA, Keith DA, Penman TD, Price OF. 2016 Resolving future fire management conflicts using multicriteria decision making. *Conserv. Biol.* **30**, 196–205. (doi:10.1111/cobi.12580)

139. Baker WL. 1992 Effects of settlement and fire suppression on landscape structure. *Ecology* **73**, 1879–1887. (doi:10.2307/1940039)

140. Keeley JE. 2006 Fire management impacts on invasive plants in the western United States. *Conserv. Biol.* **20**, 375–384. (doi:10.1111/j.1523-1739.2006.00339.x)

141. Hong R, Li J, Wang J, Zhu X, Li X, Ma C, Cao H, Wang L, Wang Q. 2023 Effects of prescribed burning on understory *Quercus* species of *Pinus yunnanensis* forest. *Front. For. Glob. Chang.* **6**. (doi:10.3389/ffgc.2023.1208682)

142. Martin KL, Hurteau MD, Hungate BA, Koch GW, North MP. 2015 Carbon tradeoffs of restoration and provision of endangered species habitat in a fire-maintained forest. *Ecosystems* **18**, 76–88. (doi:10.1007/s10021-014-9813-1)

143. Almada ED, Mendes ABV, Filho AC. 2020 Burning reasons: traditional land management using fire and environmental conflicts in Serra da Canastra National Park, Minas Gerais, Brazil. In *Socio-environmental regimes and local visions* (eds M Arce Ibarra, M Parra Vázquez, E Bello Baltazar, L Gomes de Araujo), pp. 205–224. New York, NY: Springer. (doi:10.1007/978-3-030-49767-5_10)

144. Swain DL, Abatzoglou JT, Kolden C, Shive K, Kalashnikov DA, Singh D, Smith E. 2023 Climate change is narrowing and shifting prescribed fire windows in western United States. *Commun. Earth Environ.* **4**, 1–14. (doi:10.1038/s43247-023-00993-1)

145. Sitters H, Di Stefano J, Wills T, Swan M, York A. 2018 Survey design for precise fire management conservation targets. *Ecol. Appl.* **28**, 35–45. (doi:10.1002/eap.1624)

146. Lentile LB, Holden ZA, Smith AMS, Falkowski MJ, Hudak AT, Morgan P, Lewis SA, Gessler PE, Benson NC. 2006 Remote sensing techniques to assess active fire characteristics and post-fire effects. *Int. J. Wildland Fire* **15**, 319–345. (doi:10.1071/WF05097)

147. Neary DG, Klopatek CC, DeBano LF, Ffolliott PF. 1999 Fire effects on belowground sustainability: a review and synthesis. *For. Ecol. Manag.* **122**, 51–71. (doi:10.1016/s0378-1127(99)00032-8)

148. Erdozain M, Cardil A, de-Miguel S. 2024 Fire impacts on the biology of stream ecosystems: a synthesis of current knowledge to guide future research and integrated fire management. *Glob. Chang. Biol.* **30**, e17389. (doi:10.1111/gcb.17389)

149. Espinosa del Alba C, Hjältén J, Sjögren J. 2021 Restoration strategies in boreal forests: differing field and ground layer response to ecological restoration by burning and gap cutting. *For. Ecol. Manag.* **494**, 119357. (doi:10.1016/j.foreco.2021.119357)

150. De Palma A *et al*. 2018 Challenges with inferring how land-use affects terrestrial biodiversity: study design, time, space and synthesis. In *Advances in ecological research, vol. 58, next generation biomonitoring: part 1* (eds DA Bohan, AJ Dumbrell, G Woodward, M Jackson), pp. 163–199. Oxford, UK: Academic Press. (doi:10.1016/bs.aecr.2017.12.004)

151. Reed MS, Stringer LC, Fazey I, Evely AC, Kruijsen JHJ. 2014 Five principles for the practice of knowledge exchange in environmental management. *J. Environ. Manag.* **146**, 337–345. (doi:10.1016/j.jenvman.2014.07.021)

152. Puig-Gironès R, Palmero-Iniesta M, Fernandes PM, Oliveras Menor I, Ascoli D, Kelly LT, Luke T *et al*. 2025 Supplementary material from: The use of fire to preserve biodiversity under novel fire regimes. Figshare. (doi:10.6084/m9.figshare.c.7738720)