



# OPEN Concentration gradient of plastic debris larger than 500 $\mu\text{m}$ detected across the Southwest Indian ocean

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Marine plastic pollution is increasing in the world's ocean, with the Indian Ocean understudied compared to the Pacific and Atlantic Oceans. This study investigates plastic pollution in the Southwest Indian Ocean, focusing on a size range from large debris to microplastics ( $> 500 \mu\text{m}$ ). Using visual surveys and manta trawling, we assessed plastic concentrations, compositions, and polymer types across 19 oceanographic campaigns. A total of 11,438 litter items were identified, with over 70% consisting of plastics. Larger plastic debris was predominantly observed near Glorieuses Islands during visual surveys, while microplastics were more prevalent offshore, collected through manta trawling. We observed a gradient of increasing plastic concentrations along the  $30^{\circ}/33^{\circ}\text{S}$  latitudes, from  $40^{\circ}\text{E}$  (macroplastics: 10 items/ $\text{km}^2$ ; microplastics:  $10^3$  items/ $\text{km}^2$ ) to  $65^{\circ}\text{E}$  (macroplastics:  $10^2$  items/ $\text{km}^2$ ; microplastics:  $10^5$  items/ $\text{km}^2$ ). The majority of plastic debris consisted of hard fragments, primarily polyethylene (45.7%) and polypropylene (26.7%). Our findings provide new insights into microplastic concentrations in offshore regions, highlight the significant degradation of plastic debris, and emphasize the need for further research to identify and map the Indian Ocean's garbage patch along these latitudes. **Key words:** Indian Ocean, Marine litter, Visual survey, Manta trawling, Microplastics.

Plastic pollution is accumulating in marine and coastal ecosystems<sup>1–5</sup>, causing significant socio-economic impacts and threatening wildlife worldwide<sup>6–8</sup>. Despite ongoing negotiations for a global plastics treaty aimed at ending plastic pollution, the continued mass production of plastics and inadequate waste management systems contribute to substantial emissions into the marine environment<sup>9</sup>. Marine litter enters ecosystems through various pathways, including maritime<sup>10</sup>, aerial, and terrestrial sources<sup>1,2,11</sup>, and can travel large distances via ocean currents, meanwhile breaking down into smaller fragments. There is an urgent need to identify zones of plastic accumulation and understand their long-term fate<sup>9–12</sup>.

Floating plastics, carried by wind and ocean currents, accumulate in certain areas, forming so-called “garbage patches.” While five such patches have been identified globally: two in the Pacific (North<sup>14–16</sup>; South<sup>16</sup>), two in the Atlantic (North<sup>17,18</sup>; South<sup>19–21</sup>), and one in the Southern Indian Ocean<sup>22–25</sup>, the exact location of the latter is still debated. Some studies suggest it lies in the southwest<sup>26–28</sup>, while others place it in the southeast<sup>4,23,29,30</sup> or central parts of the basin<sup>31</sup>. Numerical models predict the Indian Ocean garbage patch to be the second largest accumulation of floating plastics in the world<sup>22</sup>, as a result of countries bordering this ocean being major emitters of pollution (e.g. Southeast Asia)<sup>10,32</sup>. For example, the top 10 rivers contributing to ocean pollution are located in this ocean basin<sup>2</sup>. In these countries, waste management infrastructure is lacking, and the importation of consumable products and waste from other countries worsens the situation<sup>33</sup>. Despite this, direct observational data from this region remain limited<sup>22,23,34–36</sup>. Most research to date has focused on beach plastic debris, with evidence pointing to Southeast Asia as a primary source<sup>5,10,37–43</sup>. In contrast, offshore studies are scarce, and little

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is known about the concentration, distribution, composition, and polymer types of floating plastic pollution in the Southwest Indian Ocean. Increasing in situ surface ocean data, including both the concentration and composition of plastic debris, is essential for developing and improving plastic debris dispersion models in the Indian Ocean, as field observations provide the foundational input needed for accurate modeling.

This study aims to fill this knowledge gap by conducting extensive in situ sampling across the Southwest Indian Ocean. Our objectives are to: (i) estimate the concentration of plastic debris across the region using visual surveys and manta trawling, (ii) characterize the composition of plastic debris in terms of shape, size class, and (iii) determine the polymer types. By understanding where plastic is most likely to accumulate in this region, targeted cleanup efforts, like large-scale removal operations or technologies designed to collect plastic from the ocean, can be more efficiently deployed. Additionally, evidence of plastic accumulation hotspots is valuable for identifying the sources of pollution and informing upstream reduction strategies.

## Methods

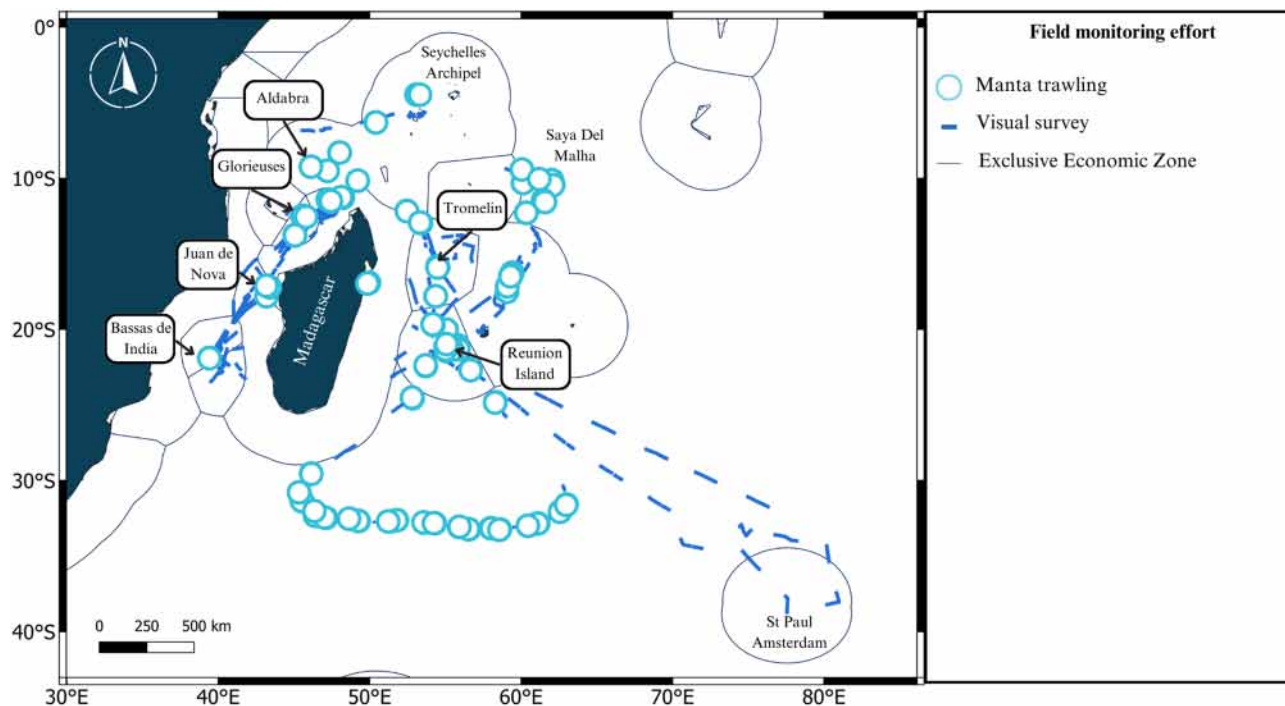
### Study site

This research utilizes data collected during 19 oceanographic campaigns that have been conducted in the Southwest Indian Ocean (SWIO, 4°/40°S – 38°/82°E, Fig. 1) between 2019 and 2023. During these opportunistic campaigns, data on the concentration of plastic debris at the ocean surface was collected using visual sea surface surveys and manta trawls in offshore sites (> 20 km from the coast) and near to remote islands (< 20 km). Campaign details, such as date, location, type of vessel, are listed in Supplementary Table S1 and Figures S1, S2.

### Visual sea surface surveys

Before each opportunistic offshore campaign, we conducted training sessions for each observer on board the vessel, focusing on the identification of floating plastic debris at the sea surface. Observations were made during daylight hours while the vessel was moving at constant speed<sup>34,44–47</sup>. At least two observers were located on the bridge to allow a visual survey area of 180 degrees. Observers used both naked eye and binoculars to make observations. There was no standard survey duration or distance, but observers recorded start/end GPS coordinates for the starting and ending points of the survey, duration of the survey, vessel speed, platform height, number of observers, and environmental parameters (sea state and cloud coverage). During these surveys, observers counted plastic debris larger than 2.5 cm up to 1 m (referred to as macroplastics<sup>48</sup> and those exceeding 1 m (referred to as megaplastics<sup>48</sup> along a 10 m wide transect. For each observation, observers noted : number of items, general type of items according to protocols<sup>48,49</sup> (e.g., clothing, rubber, ceramic, fishing gear, health materials, metal, paper, glass, wood, hard plastic, soft plastic, various and other) and sub category for each type (i.e. : soft plastic : polystyrene, more details on the Supplementary Table S2).

The plastic debris concentration was calculated with:



**Fig. 1.** Map of the Southwest Indian Ocean showing the different field trips for plastic collection at sea and on beaches. Circles correspond to manta trawl samples ( $N = 204$ ). The blue line corresponds to visual surveys ( $N = 1,884$  h effort). All information about oceanographic campaigns and programs are included in the Supplementary information Tables S1 and Figure S1.

$$Ci = \frac{Cs}{t * v * (0.01 * 2)}$$

Where  $Ci$  is the concentration plastic debris observed (item.km<sup>-2</sup>),  $Cs$  corresponds to the raw concentration of plastic debris,  $t$  is the sampling duration (h),  $v$  is the vessel speed (km.h<sup>-1</sup>) and 0.01 is the individual transect width in km<sup>14</sup>.

### Manta trawl sampling

On board, the manta net was deployed at > 30 m behind the vessel to avoid the vessel's wake to collected plastic debris at the sea surface<sup>15</sup> (all information about dimensions of nets and flowmeters are given in Supplementary Table S1). At each site, three consecutive 30-minute transects were conducted at a speed of 2 knots using an individual single-use cod-end for each transect<sup>15,50</sup>. Between transects, the net was rinsed with seawater on the outside of the net to move any missed plastic debris towards the cod-end. The cod-end was then removed, placed into an annotated Ziplock bag (date, campaign, cod-end identification number), and stored in a freezer until transferred for analysis in the laboratory ashore. For the next deployment, a new code-end was placed. For each manta trawling, the following environmental parameters were recorded: wave height (m), wind speed (m.s<sup>-1</sup>), atmospheric pressure (hPa), year, month, season (Dec-Feb: Wet season; Mar-May: Interseason 1; Jun-Aug: Dry season; Sept-Nov: Interseason 2), surface area of sampling (km<sup>-2</sup>, flowmeters, gps coordinates) and vessels characteristics. The manta net was not deployed if wave height exceeded 2 m. Once at the laboratory, each manta trawl cod-end was externally rinsed to deposit all the content on a sieve with 500 µm. Debris smaller than 500 µm was not included in the study to remove any potential contamination (e.g. microfibers).

Under light and a magnifying glass, all plastic debris were collected with ultra-fine tweezers (300 µm point diameter) and placed in a Petri dish until analysis and characterization. When all plastic debris were placed on the Petri dish, an image was taken with a camera and used to count the number of items (Nikon D7500 - lens: AF-S MICRO NIKKOR 105 mm). Then we determined for each item: (i) the shape (hard plastic: fragment, film; foam; pellet; line; non-synthetic); (ii) the dry weight (10<sup>-5</sup> g precision balance); (iii) the size class (SC1: 0.05–0.15 cm, SC2: 0.15–0.5 cm, SC3: 0.5–1.5 cm, SC4: 1.5–5 cm, ImageJ software 1.5 K<sup>51</sup>). The concentration of plastic debris was calculated by incorporating the effect of wind mixing into the calculation of the concentration of plastic debris at the sea surface<sup>52</sup>:

$$Ci = \frac{Cs}{1 - e^{-dW_b(1.5\sqrt{\frac{\rho_a}{\rho_w}C_dU^2}, k, \frac{0.96}{g}, \sigma^{\frac{3}{2}}, C_d, U^2)^{-1}}} \quad (1)$$

Where  $Ci$  is the depth-integrated concentration for the upper 5 m of the water column (item/km<sup>-2</sup>),  $Cs$  corresponds to the raw concentration of plastic debris type and size class as measured in the laboratory linked with the sampling surface area (km<sup>-2</sup>),  $d$  is the depth of the manta net,  $W_b$  is the rising velocity by plastic type and size (m.s<sup>-1</sup> determined by Lebreton et al.<sup>15</sup>,  $\rho_a$  is the air density (kg.m<sup>-3</sup>),  $\rho_w$  is the seawater density (kg.m<sup>-3</sup>),  $C_d$  is the drag coefficient (0.0012),  $U$  is the sea surface wind speed during sampling (m.s<sup>-1</sup>),  $k$  is the Karman constant (0.4),  $g$  is the gravitational constant (9.81 m.s<sup>-2</sup>) and  $\sigma$  is the wave age equal to the constant 35.

### Infrared spectroscopy ATR-FTIR

Among all items collected by manta sampling, we determined the polymer type for 1,176 items using Fourier Transform InfraRed (FTIR) spectroscopy at the UMR Softmat laboratory, in University Paul Sabatier II, France. This was done with a Thermo Nicolet Nexus 6700 instrument equipped with a diamond crystal Attenuated Total Reflection (ATR) mode and a deuterated triglycine sulphate detector. During the analysis, white background and sample spectra were obtained using 16 scans covering the wavelength range of 400 cm<sup>-1</sup> to 4,000 cm<sup>-1</sup> with a resolution of 4 cm<sup>-1</sup>. A white background spectrum was taken every 2 h to ensure accuracy. Each piece of plastic debris was pressed between the diamond crystal and the base. The diamond crystal was cleaned between each measurement to avoid any bias between spectra. The obtained spectra were corrected using the ATR thermo-correction method to obtain transmission-like spectra<sup>53</sup>. The final infrared spectra were observed using the Omnic software (version 9.9.0.473). Only spectra with more than 80% similarity to one of the spectra in the spectra database<sup>54</sup> created by the laboratory SOFTMAT at the University of Toulouse Paul Sabatier, were validated. In the cases where the similarity was less than 80%, the assignment to a specific polymer type was not made to avoid identification errors.

### Index oxidative carbonyl

To provide additional information on the level of degradation of plastic debris, a carbonyl index was determined for polyethylene (PE) and polypropylene (PP) polymer particles. These particles were matched with a similarity of over 80%. For the PE carbonyl index, the ratio between the integrated absorbance of the carbonyl peak in the range of 1,850 cm<sup>-1</sup> to 1,650 cm<sup>-1</sup> and the methylene scissoring peak in the range of 1,500 cm<sup>-1</sup> to 1,420 cm<sup>-1</sup> was calculated. The Specified Area Under these two bands (SAUB) provided by Almond et al.<sup>55</sup> was calculated using Omnic software with the options band analysis tool. For the PP carbonyl index, the ratio between the peak height at 1,715 cm<sup>-1</sup> and the area under the band in the range of 1,500 cm<sup>-1</sup> to 1,450 cm<sup>-1</sup> was also calculated. The area under the band and peak height were calculated by the same software using the options peak analysis tool. For each measurement, a flat baseline was applied using the data between 4,000 cm<sup>-1</sup> and 2,000 cm<sup>-1</sup>.

## Statistical analysis

Before each test, normality and homoscedasticity were assessed using the Shapiro and Levene tests, respectively, for each quantitative variable of item concentration. None of our data follow a normal distribution. For each method (visual sea surface surveys and manta net), the quantity of items, expressed as an integer percentage (%), was modeled based on explanatory variables. For visual observation, only the category was considered, while for the Manta net, the variables category, polymer, and size were used. These models were fitted using a generalized linear model (GLM) with a Poisson distribution.

Regarding total concentration, after visualizing our data using Q-Q plots and the Shapiro-Wilk test, we found that our data was not normally distributed (Script on the supplementary data Figure S3). Therefore, we decided to examine the influence of dependent qualitative explanatory variables ( $X_n$ ) including continuous (latitude, longitude) and categorical (season, platform height [m: 0–3, 3–6, 6–9, 9–12], cloud coverage [%: 0, 25, 50, 75, 100], and sea state), on the quantitative variable ‘abundance of plastic debris’ observed from visual survey effort using a generalized additive model (GAM) with a Poisson family. A GAM with a Poisson family was also used to explain the concentration of plastic debris collected from manta trawl sampling based on the following explanatory variables: latitude, longitude seasons (Dec-Feb: Wet; Mar-May: Interseason1; Jun-Aug: Dry; Sep-Nov: Interseason2) and distance (offshore, onshore).

For each GAM, we selected models with Akaike’s Information Criteria adjusted for small sample sizes (AICc). All statistical tests were conducted in the R computing environment (R Core Team, 2023).

## Results

### Overall

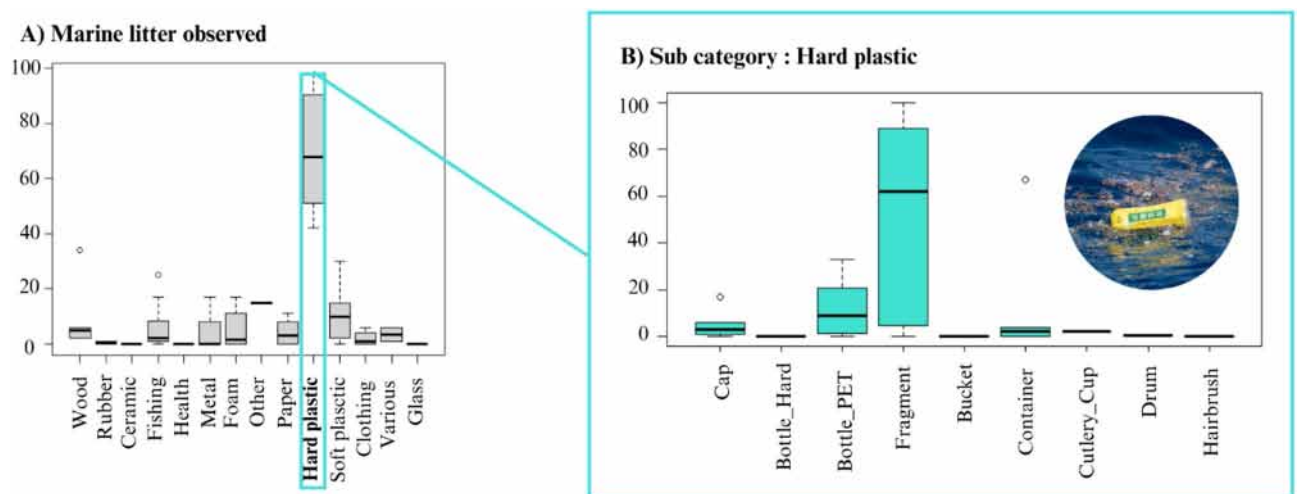
A total of 11,438 marine debris items were recorded, from sea surface survey (75.7%) and manta trawling (24.3%).

### Visual survey composition

During the 1,884 h of sea surface survey, 8,655 pieces of marine litter were observed, of which 97% were plastic debris. Hard plastic was the most frequently recorded marine debris category in all surveys with averages of 70% (Fig. 2.A), followed by soft plastic (film, sheet), fishing gear (line, rope, bucket, drum) and foam, while glass, ceramic, and healthcare-related items were generally less frequently observed (GLM, Table S3.1). The category, “Hard Plastics” was predominantly composed of “fragment” with 64% for floating plastic debris (Fig. 2.B), followed by “PET bottles” and containers with 15% and 6% of observations respectively (GLM, Table S3.2).

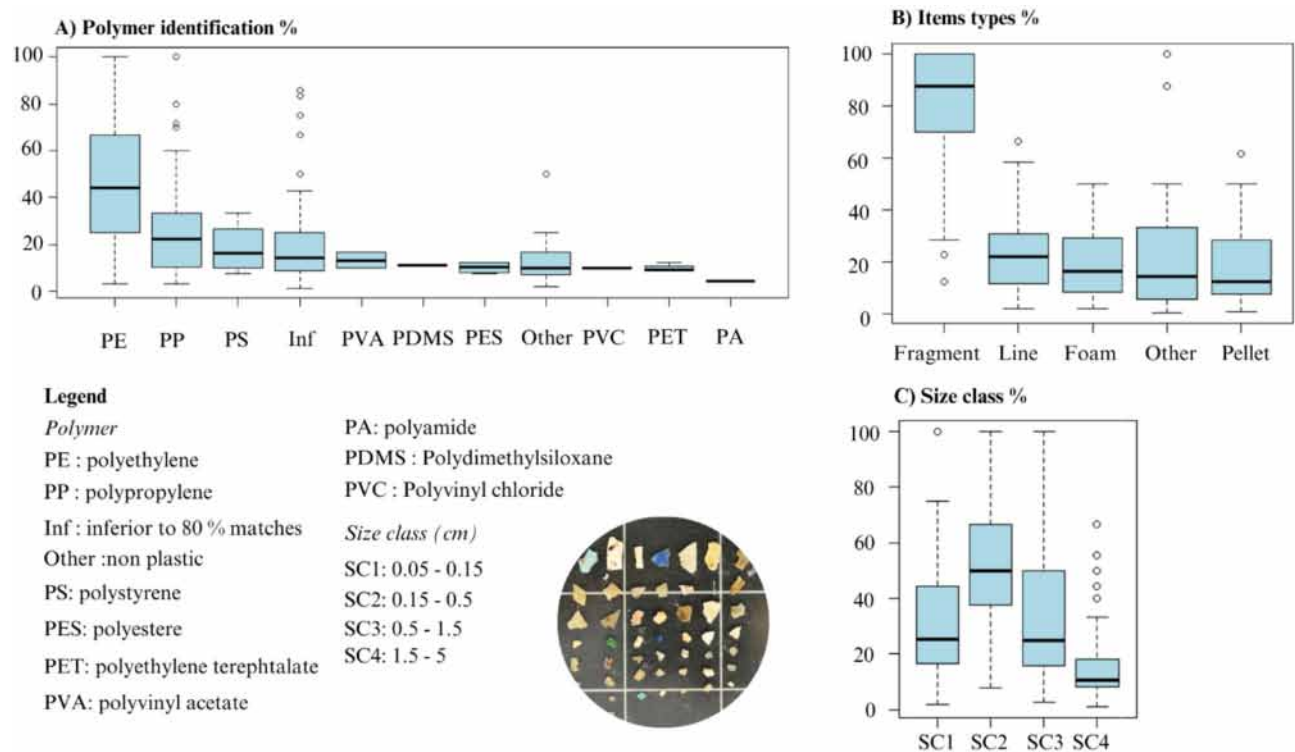
### Manta sampling composition

A total of 2,745 items were collected by manta trawl, among them 43% ( $N = 1,176$ ) where analyzed using ATR-FTIR. A large proportion of this items, 84% ( $N = 985$ ) were successfully matched to the database with a score  $> 80\%$ , while 16% ( $N = 191$ ) remained unidentified. Items that were successfully identified were composed of hard plastic ( $81.2 \pm 1.51\%$ ), non-synthetic items ( $25.9 \pm 5.99\%$ ), line ( $23.8 \pm 1.57\%$ ), foam ( $18.6 \pm \%$ ) and pellet ( $19.6 \pm 1.51\%$ , Fig. 3.A, GLM, Table S3.3). These microplastics were predominantly composed of PE (45.7%) followed by PP (26.7%) of various shapes and size classes (Fig. 3.B, Table 1, GLM, Table S3.4). The size classes SC2 (0.15–0.5 cm) has been mainly collected (Fig. 3.C, Table 1; GLM, Table S3.5). Foam-shaped plastics exhibited a diverse range of polymers. Plastic debris found floating inshore ( $< 12$  miles,  $N = 220$ ) consisted of 40% PE, 34% PP, and 1.8% (PVA, PS, PDMS), whereas offshore ( $> 12$  miles,  $N = 956$ ) debris was composed of 62% PE, and 19% PP and 1.5% (PVA, PES, PS, PA, PET). There was no significant difference in polymer composition with distance from the coastline. PE and PP carbonyl index showed no significant difference between onshore and offshore



**Fig. 2.** Composition of marine litter observed from visual survey (A) Percentage abundance of marine litter by category, (B) percentage abundance of hard plastic subcategory.





**Fig. 3.** Percentage abundance (%) of items collected by manta net by (A) polymers, (B) types and (C) size classes.

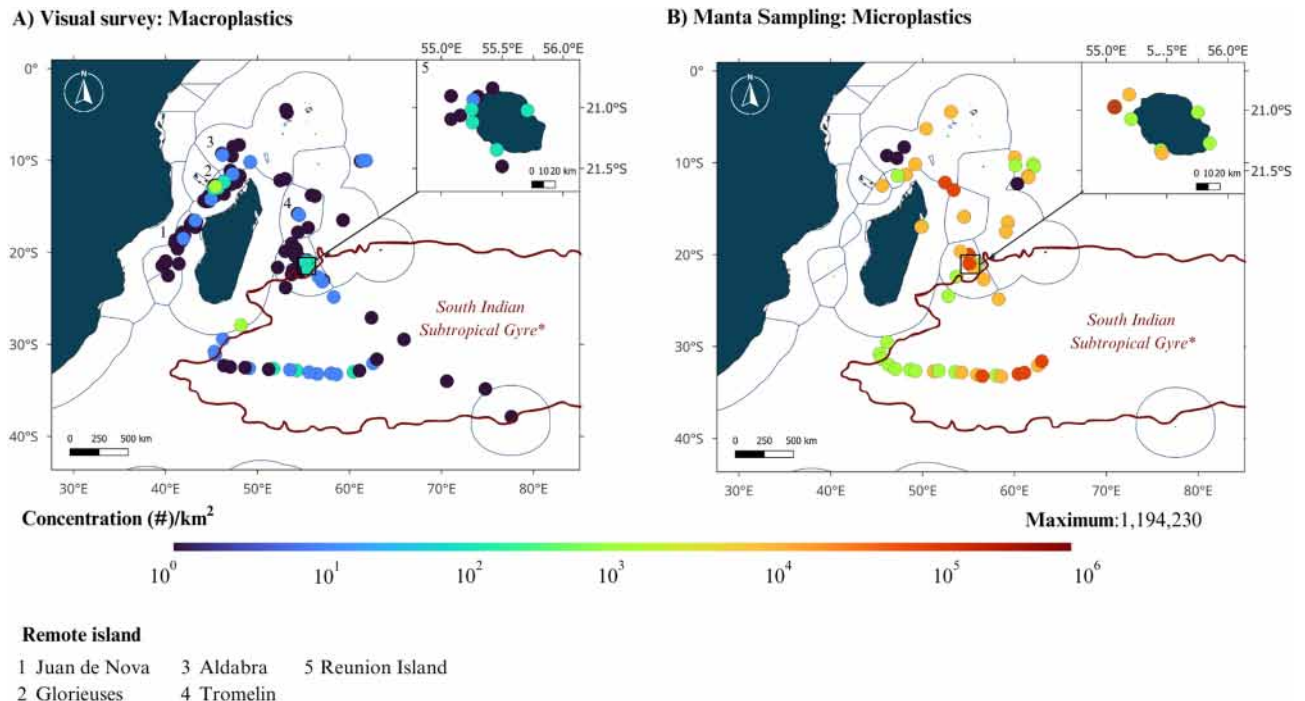
		Size class (cm)	Type H	Type N	Type F	Type P	All
A)	All manta sampling	Small Microplastic (SC1: 0.05–0–0.15)	2,131 ± 380 (56,163)	50 ± 16 (1,841)	64 ± 19 (1,841)	-	2,246 ± 386 (56,943)
		Large Microplastic (SC2: 0.15–0.5)	4,285 ± 885 (171,610)	472 ± 153 (28,747)	135 ± 40 (5,525)	10 ± 6.2 (920)	4,903 ± 957 (180,971)
		Mesoplastic (SC3: 0.5–1.5)	1,536 ± 223 (29,468)	537 ± 108 (16,575)	28 ± 11 (1,841)	10 ± 6.2 (920)	2,112 ± 293 (46,043)
		Small Macroplastic (SC4: 1.5–5)	231 ± 52 (8,287)	281 ± 56 (7,367)	2.7 ± 2.7 (568)	-	596 ± 95 (15,654)
B)	Onshore	Small Microplastic (SC1: 0.05–0–0.15)	2,710 ± 998 (56,163)	92 ± 47 (1,841)	85 ± 40 (1,841)	-	2,889 ± 1,013 (56,943)
		Large Microplastic (SC2: 0.15–0.5)	6,553 ± 3,005 (171,610)	557 ± 203 (7,800)	216 ± 102 (5,525)	-	7,327 ± 3,180 (180,971)
		Mesoplastic (SC3: 0.5–1.5)	1,408 ± 195 (16,575)	602 ± 195 (9,433)	11 ± 11 (668)	-	2,022 ± 438 (16,575)
		Small Macroplastic (SC4: 1.5–5)	266 ± 87 (3,703)	218 ± 65 (2,222)	-	-	484 ± 128 (4,444)
C)	Offshore	Small Microplastic (SC1: 0.05–0–0.15)	1,904 ± 358 (36,613)	34 ± 13 (1,060)	56 ± 21 (1,841)	-	1,994 ± 363 (39,613)
		Large Microplastic (SC2: 0.15–0.5)	3,397 ± 361 (23,746)	438 ± 199 (28,747)	103 ± 39 (3,683)	14 ± 8.7 (920)	3,954 ± 471 (47,885)
		Mesoplastic (SC3: 0.5–1.5)	1,586 ± 274 (29,468)	512 ± 129 (16,575)	34 ± 15 (1,841)	14 ± 8.7 (920)	4,518 ± 371 (46,043)
		Small Macroplastic (SC4: 1.5–5)	217 ± 65 (8,287)	307 ± 74 (7,367)	3.7 ± 3.7 (558)	-	528 ± 122 (15,654)

**Table 1.** Mean + SE (max value) of numerical concentration per Kilometer square (items.km<sup>2</sup>) by plastic type and size through a) the 204 manta sampling in the South West Indian ocean, B) plastic debris collected onshore and C) plastic debris collected offshore. Plastic type H include pieces of hard plastic, plastic sheet and film, type N encompasses plastic lines, ropes and fishing nets, type F are pieces made of foamed material and type P are pre-production plastic pellets.

samples. Nevertheless, there was a noticeable trend indicating increased degradation of PP with distance from the coast towards offshore (Supplementary Figure S4, Table S4).

#### Concentration of plastic debris from macro to micro

We observed a mean (± sd) macroplastics concentration of 90.8 (± 411) items.km<sup>-2</sup> (Fig. 4.A). The maximum concentration, reaching 4,585 items.km<sup>-2</sup>, was observed around Glorieuses Islands in the north of the Mozambique Channel. A gradient of macroplastics was also observed from the west (10<sup>0</sup> items.km<sup>-2</sup>) to the east (10<sup>2</sup> items.km<sup>-2</sup>) on the latitude 30/33°S. The average abundance at 32°S latitudes was 40 (± 869) items.km<sup>-2</sup> (max: 237 items.km<sup>-2</sup>, 32°S and 60°E). The GAM confirmed the influence of the wet season and the interseasonal period in October, which had an influence on the increase in observed macroplastic concentration,



**Fig. 4.** The concentration of plastic debris by abundance recorded from (A) manta trawl sampling (size class: 500  $\mu\text{m}$  – 5 mm) and (B) visual surveys (size class: 2.5–100 cm). \*The South Indian Subtropical Gyre boundary is adapted from Lebreton<sup>2</sup>.

as well as good weather conditions characterized by little or no cloud coverage. However, the higher the vessel platform height, the lower the concentration of smaller sizes of plastic debris was observed. The longitude and latitude variables also show significant relationships, with significant smooth terms ( $p < 2e-16$ ,  $n = 178$ ). The model explains 65.9% of the deviance, with an adjusted  $R^2$  of 0.72, indicating a good fit of the model to the data. (Figure S5.1)

Regarding microplastics, the highest concentration was collected on the northwest of Reunion Island from the IOTA campaigns with  $10^7$  items.km<sup>-2</sup>. The lowest was recorded around Aldabra with  $10^0$  items.km<sup>-2</sup> (Fig. 4.B). Overall, the microplastic concentration exhibits a noticeable increasing longitudinal gradient along latitude 30°/33°S, ranging from  $10^3$  items.km<sup>-2</sup> at 40°E to  $10^5$  items.km<sup>-2</sup> at 65°E. The GAM revealed significant effects of several factors on the variable microplastics concentration. The seasonal effects indicate that both the interseason 1 and interseason 2 periods, as well as the wet season, were associated with a significant decrease in microplastics concentration, as well as the distance from the shore ( $p < 2e-16$ ,  $n = 206$ , ScriptR results from GAM test, Supplementary Figure S5.2). The smooth terms for longitude and latitude were highly significant, indicating that both geographical coordinates are important predictors of macroplastic concentration ( $p < 2e-16$ ,  $n = 206$  for ScriptR results from GAM test, Supplementary Figure S5.2). The model explained 46.3% of the deviance, with an adjusted  $R^2$  of 0.416, suggesting a reasonable fit to the data.

## Discussion

### The offshore concentration gradient

Our study confirms the presence of sea surface plastic pollution across the Southwest Indian Ocean. We observed a gradient of increasing plastic concentrations along the 30°/33°S latitudes, from 40°E (macroplastics: 10 items.km<sup>-2</sup>; microplastics:  $10^3$  items.km<sup>-2</sup>) to 65°E (macroplastics:  $10^2$  items.km<sup>-2</sup>; microplastics:  $10^5$  items.km<sup>-2</sup>). In this region, the macroplastic concentration observed at the 65°E is in the same order of magnitude as previous research<sup>34</sup>. In the North Pacific<sup>15</sup>, macroplastics (> 5 cm) concentrations range from 10 to  $10^3$  items.km<sup>-2</sup>. In other oceans, data is reported for macroplastic concentrations<sup>22,28,56</sup>, but the classification of size categories is inconsistent or not standardized. Most authors adopt a threshold of 0.5 cm, considering anything larger than microplastics as macroplastics, while others use > 2.5 cm or > 5 cm. Although macroplastics are a major source of secondary microplastics by degradation, their monitoring in the ocean remains limited to mostly visual observations that may underestimate concentrations<sup>37</sup>, while larger net device may be more appropriate to study their distribution yet at this stage only data from the North Pacific was reported using such methods<sup>25</sup>.

Microplastics surface concentrations observed in the Indian Ocean are higher than those observed in the South Atlantic ( $10^3$  items.km<sup>-2</sup>)<sup>21</sup> or in the South Pacific ( $10^4$  items.km<sup>-2</sup>)<sup>16</sup>, in similar orders of magnitude of maximum concentrations reported in the North Atlantic ( $10^5$  items.km<sup>-2</sup>)<sup>58</sup> but lower than those in the North Pacific (>  $10^6$  items.km<sup>-2</sup>)<sup>58</sup>. Regional studies in the Southern Indian Ocean<sup>22,23,34</sup> and Bay of Bengal<sup>59</sup> have reported concentrations ranging from  $10^2$  to  $10^6$  items.km<sup>-2</sup>, however, differences in sampling methods and mesh sizes (ranging from 0.02 to 0.09 cm) make comparisons challenging, highlighting the value of regional

scientific expeditions for gaining robust information on this issue. As we identified a concentration gradient between 30° and 33°S, further field work is needed to continue sampling along this gradient further east and accurately delineate the accumulation area.

### Inshore concentrations

High concentrations of macroplastics ( $10^3$  items.km<sup>-2</sup>), primarily composed of fragmented hard plastics, followed by caps, water bottles, and flip-flops, were observed in the northern Mozambique Channel, particularly around the Glorieuses Islands, while microplastics were more prevalent farther offshore. Studies identified water bottles as the predominant sources of pollution close to urbanized areas (e.g., Mayotte<sup>60</sup>, Seychelles<sup>10,37</sup>, Madagascar<sup>43</sup>, Reunion Island<sup>42</sup>, while flip-flop and fishing-related debris were more prevalent in more remote locations (e.g., Saint Brandon<sup>38</sup>, Tromelin, Juan de Nova<sup>61</sup>, Glorieuses, Europa, Sainte-Marie<sup>43</sup>). However, by modeling sea currents and identifying sources of plastic debris beached on remote islands by the Brand Audit method in the region<sup>62</sup>, studies have identified Southeast Asia as a major origin of plastic bottles<sup>32,37,43,63,64</sup>. These macroplastics are likely driven across the Indian Ocean by the South Equatorial Current<sup>65</sup>, and can degrade into microplastics through exposure to sunlight and wind. They may then accumulate in mesogyre, such as those in the Mozambique Channel, or be trapped by mesoscale anticyclonic eddies generated by island mass effects<sup>66,67</sup>, like the one northwest of Reunion Island, potentially explaining the extremely high microplastics concentrations we observed  $10^7$  items.km<sup>-2</sup>. The threat of plastic pollution is not limited to the ocean surface, it can also travel long distances and wash up on remote coastal regions with unique ecosystems that are considered biodiversity hotspots<sup>61,68</sup>.

### Consequences of this plastic pollution in the region

Depending on the size of plastic debris (from macro to microplastics) and their location (inshore or offshore), they can impact biodiversity in various ways. In our study region, the Chagos Archipelago<sup>69</sup>, Reunion Island<sup>70</sup>, Mayotte<sup>70</sup>, or the Eparses Islands<sup>71</sup> are all defined as hotspots of biodiversity, with the hosting of marine megafauna like sea turtles<sup>72</sup>, seabirds<sup>73</sup>, and cetaceans<sup>74,75</sup>. Previous studies in the region have reported entanglement of post hatchling sea turtles on nesting beaches during emergence<sup>76,77,69</sup>, as well as ingestion by sea turtles<sup>43,78</sup> and seabirds<sup>79</sup> in pelagic environment. A study on Loggerhead sea turtle by-catch in the south west Indian ocean, indicated that ingested plastic debris was primarily composed of plastic bottle caps<sup>43</sup> that were likely originating from Southeast Asia. This finding agrees with visual surveys of floating macroplastics reported here. Two other studies have reported high ingestion risk for seabirds, specifically petrels, around latitudes 30/33° S in the Indian Ocean<sup>80,81</sup>. Adding to risks of ingestion, plastics can adsorb toxic substances<sup>82</sup>, as well as pathogenic bacteria<sup>42</sup>, which can cause diseases in tropical ecosystems<sup>83,84</sup> such as those found in the Southwest Indian Ocean.

### Limitation and perspective

From 1988<sup>86</sup> to 2024, studies in the Indian Ocean have investigated floating plastic debris, using either visual surveys of macroplastics or in situ sampling of microplastics, with various collection protocols and size classification methods<sup>64</sup>. In our study, we collaborated with research groups from multiple countries or regions (Madagascar, Seychelles, Reunion Island, Eparses Islands) and organizations, aiming to harmonize protocols and size classes as much as possible. Naturally, some differences remained, such as the number of observers on board, the size of the nets used, and the sampling periods (e.g., during or outside the rainy season). All relevant details are provided in the supplementary data. For future research, it would be essential to establish a standardized nomenclature and clearly define size classes, and make raw data files openly available<sup>48,62</sup>. Concerning the concentrations observed around remote islands (e.g., the Eparses Islands), it would be valuable to conduct seasonal monitoring of beached marine litter and determine its origin. Similarly, at latitudes 30°–33° South, further research is needed to define the extent of the Indian Ocean plastic patch and to precisely locate it, particularly to assess whether higher concentrations may be found further East. A large quantity of water bottle caps was identified during our surveys, it is difficult to know whether these come from bottles that were originally discarded on land or at sea, from shipping or fishing vessels. As it could well be a result of both, it is therefore recommended to improve access to potable water in countries neighboring the Indian Ocean to reduce the consumption and discard of bottled water, as well as enforcing stricter controls of waste management aboard vessels crossing the region. These actions could be integrated into the United Nations international treaty on plastic pollution to end plastic pollution, including in the marine environment.

### Conclusion

Our sampling represents the largest geographical survey and robust sampling of floating macro and microplastics in the Southwest Indian Ocean region to date. This provides valuable in situ data for improving the understanding of plastic particle distribution and the location of the floating plastic accumulation in the region. As we identified a concentration gradient at latitude 30/33°S, we suggest conducting further research toward the central and eastern parts of the Indian Ocean basin to investigate and potentially delineate the presence of plastic patch. These results offer a first step in understanding plastic dispersion in the region, helping to guide future research, cleanup efforts, and strategy development to reduce pollution before it reaches the ocean and eventually degrades into smaller particles.

### Data availability

The datasets generated during the sea surface survey and manta trawling are available at Figshare in the [Raw data of plastic debris collected in the Southwest Indian Ocean by manta trawling] repository by DOI: 10.6084/

m9.figshare.26828164 and [Raw data of plastic debris collected in the Southwest Indian Ocean by visual survey respectively] by DOI: 10.6084/m9.figshare.28451912.

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## Author contributions

M.T. and L.L. designed the study; M.T., A.F., A.R., A.S., G.F., M.C., P.M., M.A., S.J., P.J., A.B., S.J.R., M.E., A.T.H., M.L.C., L.L. managed expeditions cruises or beach monitoring or directly facilitated the expeditions; M.T., A.F., A.R., G.F., V.M., M.C., T.M., S.J. collected the samples; M.T., A.R., M.C., V.M., J.G., analyzed the samples at laboratory; M.T. conducted the data analyses and the calculations, and prepared figures and tables; M.T. wrote the manuscript. All authors reviewed and edited the manuscript.

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## Declarations

## Competing interests

M.T., L.L., M.E., S.J.R. are employed by The Ocean Cleanup, a non-for-profit developing and scaling technologies to rid the oceans of plastics, headquartered in the Netherlands. A.F., A.R., A.S., G.F., M.C., P.M., A.B., M.A., S.J., T.M., P.J., A.T.H., M.L.C. declare no competing interests. All the remaining authors declare no competing interests.

## Additional information

**Supplementary Information** The online version contains supplementary material available at <https://doi.org/10.1038/s41598-025-02893-0>.

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